

## Original Research Article

# Multi-way Degradation and Process Optimization of Phenol from Simulated Wastewater System

### ABSTRACT

This research was based on the comparative study between microbial, enzymatic and photocatalytic phenol degradation. Different experiments were carried out under three distinct methodologies ~~and that~~ seeked to examine which method is more feasible between them through various aspects. For the microbial study, *E-coli* was used for phenol degradation at an optimum condition of *E-coli*. In ~~an-the~~ enzymatic study, peroxidase was extracted from soybean seed hulls, and it was purified. The purified peroxidase enzyme was applied in phenolic solution at neutral pH. The H<sub>2</sub>O<sub>2</sub>/UV/TiO<sub>2</sub> scheme was adopted in the photocatalytic treatment of phenol. Maximum phenol degradation was observed in photocatalysis. From this comparative study, a microbial method was found to be more time consuming and an enzymatic method having-require more steps to perform the experiment performed while photocatalysis had-took less time with a more feasible methodresults.

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*Keywords: Comparative study, microbial treatment, enzymatic treatment, photocatalytic treatment, first-order reaction kinetics*

### 1. INTRODUCTION

Recently, considerable attention has been received by many researchers on biodegradation of aromatic compounds ~~by many researchers~~ due to their toxicity. Among them, phenol and its derivatives are a standard compound in wastewater of many industries such as oil refineries [1], coal refining, petroleum, textiles and pharmaceuticals [2]. It is quite known ~~related-to-that~~ the toxicity of phenols towards the whole environment is high and thus has been incorporated in the list of pollutants by the U.S. Environmental Protection Agency [3]. Many researchers are engaged in research on phenol degradation by diverse techniques and methods. The attention is that to investigate which technology will be most feasible, eco-friendly, cost-effective and time ~~abstaining-saving-and-this-idea~~ is the primary goal of the present investigation. The present study ~~comprise~~ are three ~~parts-methods~~ viz. microbial degradation, enzymatic degradation and photocatalytic degradation.

Until today, many investigators have ~~been~~ reported numerous types of microorganisms ~~to~~ that remove phenol from wastewater. From the literature review ed, some microorganisms can consume phenol as a sole source of carbon and energy. These bacterial species include *Streptococcus epidermis* [4], *Escherichia coli*, *Micrococcus sp.*, *Brucella sp.* [5], *Bacillus subtilis*, *Pseudomonas putida*, *Acinetobacter calcoaceticus*, *Bacillus subtilis* [6-8] and *Streptococcus sp.* [8].

Besides, enzymes are applied in biodegradation study of the phenol. Enzymes play a vital role in phenol biodegradation reactions as a biocatalyst. These enzymes include Peroxidase, Chloroperoxidase, Lignin peroxidase, Mn-peroxidase [9] and catalase [10] that isolated from specific plants viz. soybean [11], horseradish, radish [12], and their materials such as seeds

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40 [13], leaves [14], stem [15], roots [16]. Tyrosinase and Laccase [9] are obtained from  
41 different fungal species.

42  
43 In recent years, photocatalysis has been developed in wastewater treatment. In this  
44 technique, some photocatalysts and their chemically modified transformations were  
45 employed for the photodegradation of toxic compounds. The TiO<sub>2</sub> and ZnO were broadly  
46 ~~worked-tested~~ as ~~a~~-photocatalysts ~~used~~ in this technique [17-20]. Many researchers increase  
47 the efficiency of a catalyst by doping with metals such as Ag, Fe, Pr, Co, V under various  
48 illumination systems [21]. Some researchers synthesized bimetallic or trimetallic  
49 transformations for degradation study [22].

50  
51 Here, we focus on all related aspects or parameters to select a better, efficient, cost-effective  
52 and feasible degradation technique. From the overall primary study, we use *E. coli* for the  
53 microbial study while peroxidase extracted from soybean seed hulls and selected for the  
54 further process of phenol degradation. Alike we introduced single TiO<sub>2</sub> nanoparticles in  
55 phenolic wastewater under both UV and Solar light.

## 56 2. MATERIAL AND METHODS / EXPERIMENTAL DETAILS / METHODOLOGY

### 57 2.1 Materials

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59 All analytical grade and HPLC grade chemicals were purchased from Fisher scientific and  
60 Himedia, Mumbai, India. Milli-Q water used for chemical preparations obtained from Milli-Q  
61 make of Shimadzu, Japan. *E-coli* microbial culture ~~gave-was given~~ by my friend. Soybean  
62 seeds were collected from agricultural fields and washed thoroughly with distilled water.

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### 63 2.2 Microbial Methodology

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67 *E-coli* bacterial culture was grown on slants of nutrient agar medium for further microbial  
68 phenol degradation study and stored at 4°C until further use. Then the minimal salt medium  
69 was prepared as Na<sub>2</sub>HPO<sub>4</sub> (33.9 g), KH<sub>2</sub>PO<sub>4</sub> (15 g), NH<sub>4</sub>Cl (5 g), NaCl (2.5 g), 2 ml of  
70 MgSO<sub>4</sub> (0.1 M) and 0.1 ml of CaCl<sub>2</sub> (1 M) per liter for actual degrading study [4]. All media  
71 and required glassware autoclaved at 121°C and 15 lbs for 15 min. for sterilizing before the  
72 commencement of experiments. Four consecutive same interval different concentrations of  
73 phenolic wastewater were prepared in the range between 250 mg/L to 1000 mg/L in  
74 phosphate buffer with pH 7.0. The reaction mixture ~~had-contain~~~~ed~~~~ing~~ only MSM media and  
75 phenol that was used as a control mixture in ~~a-the~~ microbial study. Similarly, bacterial  
76 inoculum ~~had-been-was~~ added to the control mixture for further phenol degradation study.  
77 Experiments were carried out in a 250 ml conical flask containing 50 ml of MSM media with  
78 phenol concentration of above-given range. The mixture was incubated at room temperature  
79 (37°C ± 2) on the shaker (100 rpm). Samples were collected ~~and tested~~ at every 24 h time  
80 interval for five days.

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81  
82 The samples were centrifuged, and the remaining phenol concentration ~~had-been~~  
83 determined quantitatively by direct UV-visible spectrophotometric method [23]. Optical  
84 density was measured at λ<sub>max</sub> = 269 nm. Remaining concentration of phenol (%) was  
85 calculated ~~as-using~~ following formula:

$$\% \text{ Phenol degradation} = \frac{\text{Absorbance of sample}}{\text{Slope phenol degradation (by graph)}} \dots \dots \dots (1)$$

### 86 2.3 Enzymatic Methodology

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89 | The experimental procedures of SBP extraction and purification ~~were~~ followed with some  
90 | modifications reported by Liu *et al.*, 2005. The fresh soybean seed hulls was weighed and  
91 | washed with milli-Q water. These cleaned seeds were soaked in milli-Q water ~~for~~ overnight.  
92 | The soaked seeds were smashed and blended with 500 ml milli-Q water for 10 to 15 min.  
93 | Then the homogenized mixture was filtered through cheesecloth and after that filtrate of  
94 | cheesecloth centrifuged at 10,000 rpm for 20 min at 4°C. The collected supernatant was rich  
95 | in proteins.

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97 | The SBP purification process was performed as reported in Liu *et al.*, 2005. The process  
98 | included three steps. ~~A-The~~ first step was acetone-ammonium sulphate cooperation  
99 | precipitation. It comprised both acetone and ammonium sulphate precipitation  
100 | simultaneously. The volume of acetone taken 0.3 fold of the original amount and solid  
101 | ammonium sulphate added to form up to 45% saturation. This combination was placed in a  
102 | refrigerator for 2 h. After that, the mixture was centrifuged for 15 min at 5000 to 7000 rpm.  
103 | The supernatant and precipitant were collected separately. This 45% saturation was  
104 | continued to 75% saturation by adding solid ammonium sulphate again with 0.3 fold acetone  
105 | in the supernatant. ~~A-The~~ mixture was centrifuged for 15 min at 5000 to 7000 rpm. Only one  
106 | condition followed that the acetone was pre-stored in a refrigerator and that cooled acetone  
107 | was added under a cold atmosphere in all our experimental sets. The resulted precipitants  
108 | were dissolved in milli-Q water to get primary purified SBP. The second step consisted of  
109 | acetone precipitation ~~alone~~. The volume of acetone mixed as 1.4 fold separately into  
110 | the primary purified SBP. ~~A-The~~ mixture was centrifuged for 15 min at 5000 to 7000 rpm.  
111 | The resulted precipitant was dissolved in milli-Q water to get secondary purified SBP. The  
112 | third step included only zinc sulphate precipitation. Before introducing zinc sulphate into the  
113 | enzyme solutions, the pH was adjusted ~~on-to~~ eight by HCl or NaOH and then 1.0mol L<sup>-1</sup> zinc  
114 | sulphate solution was mixed to form 0.015 mol/l zinc concentration. ~~A-The~~ mixture was  
115 | centrifuged for 15 min at 5000 to 7000 rpm. Lastly, the supernatant was collected and  
116 | denoted as highly purified SBP enzyme solution [24].

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117 | Enzyme assay and protein content were examined after each purification step by the  
118 | procedures ~~in-described by~~ Kolhe *et al.*, 2015 [13]. The RZ values were assayed after each  
119 | purification steps. The purified SBP stored at 4°C untill the further use of ~~an-the~~ enzyme.  
120 | Different concentrations of phenolic wastewater were prepared in the range between 250  
121 | mg/L to 1000 mg/L in phosphate buffer with pH 7.0. The reaction mixture contained 50 ml  
122 | phenolic wastewater, 30 per cent H<sub>2</sub>O<sub>2</sub> and enzyme solution. The sample was collected as a  
123 | control before kept for reaction and analyzed it. This combination kept on a rotary shaker for  
124 | 10 h, and aliquots were collected at every 1h time interval.

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125 | The remaining phenol concentration of each sample ~~had-was~~ determined quantitatively by  
127 | the direct UV-visible spectrophotometric method at phenol  $\lambda_{max}$ . The remaining concentration  
128 | of phenol (%) was calculated by formula 1.

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## 131 | 2.4 Photocatalytic Methodology

132 | The third methodology opted ~~as-for~~ photocatalytic degradation of phenol. In this study, TiO<sub>2</sub>  
133 | nanoparticles were used as ~~a-the~~ photocatalyst while 11 watts of UV lamp was used as  
134 | illumination for energy. Various concentrations of phenolic wastewater were prepared in the  
135 | range between 250 mg/L to 1000 mg/L. The pH range kept as 2, 4, 6, 8 and 10 and adjusted  
136 | with 0.1 M HCl and 0.1 M NaOH solutions. The retention time was 10 h, but samples were  
137 | collected at every 1h time interval. The reaction mixture contained 50 ml phenolic solution,  
138 | 30% H<sub>2</sub>O<sub>2</sub> and TiO<sub>2</sub> nanoparticles. The sample was taken as a control before kept on a  
139 | magnetic stirrer for reaction and analyzed it.

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142 | The remaining phenol concentration of each sample ~~had was~~ determined quantitatively by  
143 | the direct UV-visible spectrophotometric method at phenol  $\lambda_{max}$ . The residual concentration  
144 | of phenol (%) was calculated by formula 1. The first and second order kinetics study were  
145 | evaluated from graphs of log concentration versus irradiation time [25].  
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### 147 | 3. RESULTS AND DISCUSSION

#### 148 | 3.1 Microbial Treatment

151 | The phenol degradation performance of *E. coli* strain was examined for different phenol  
152 | concentrations viz. 250 mg/L, 500 mg/L, 750 mg/L and 1000 mg/L at various time intervals.  
153 | The per cent phenol degradation was derived based on residual phenol concentration.  
154 | Figure 1 ~~shows the~~ effect of phenol concentration ~~indicating shows~~ that 60.07% phenol  
155 | degradation ~~was~~ observed at 250 mg/L phenolic concentration at neutral pH after 96 h. ~~As~~  
156 | the phenolic concentration increases the phenol degradation decreases. Hence, only  
157 | 11.75% phenol degradation ~~was~~ observed ~~in at~~ 1000 mg/L phenolic concentration at neutral  
158 | pH after 96 h. Reshma *et al.* ~~also gave used E. coli a treatment of E. coli on~~ phenolic  
159 | wastewater. They obtained 100% phenol degradation for 10 mg/L phenolic solution. We had  
160 | only 60.07% phenol degradation because 250 mg/L concentration was much more than 10  
161 | mg/L concentration. Some bacterial strain may have died at ~~a more this high~~ phenolic  
162 | concentration; hence, the *E. coli* bacterial strain ~~had did~~ not achieved 100% phenol  
163 | degradation.  
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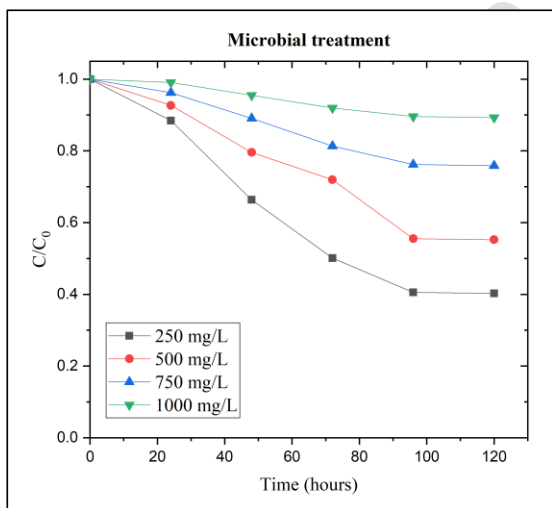
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167 | Fig. 1. Phenol degradation by microbial treatment for different concentrations of the  
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#### 170 | 3.2 Enzymatic Treatment

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The SBP was extracted from soybean seed hulls by blending it for 10 to 15 min. During the blending of soybean seed hulls, the blended material was lightly warmed, but this ~~thing~~ ~~is was~~ not essential because the SBP activity persisted up to 75°C [11]. A volume of the original enzyme solution was recorded as 530 ml. Table 1 shows the enzyme purification

176 steps and their characteristics. A product of the last purification step having 71.01% recovery  
 177 and 1.12 RZ value which is near about 1.32 RZ value reported in Liu *et al.* [24]. This enzyme  
 178 purification method is more comfortable and cost-effective than other purification methods  
 179 because it is merely based on only precipitation technique. Total volume, total activity,%  
 180 recovery, protein content, specific activity, fold purification and RZ value for each step **were**  
 181 **are shown** in table 1.

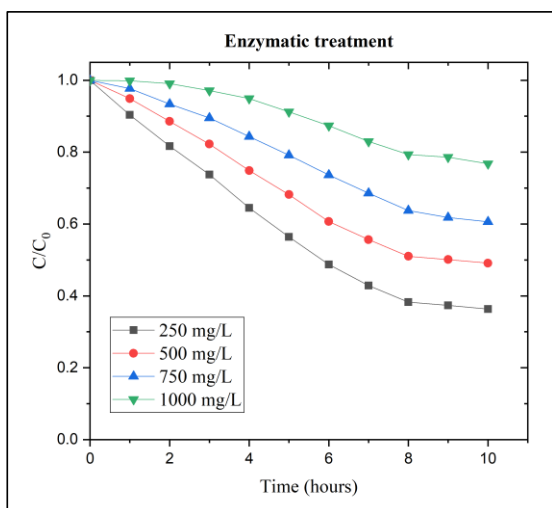
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182  
 183 **Table 1. Purification steps and their characterization of SBP**

Steps	Total Volume (ml)	Total Activity (U/ml)	Recovery (%)	Protein Content (mg/ml)	Specific Activity (U/mg)	Fold Purification	RZ value
Original enzyme solution	530	6.091	100	2.325	2.62	1	0.19
Acetone-ammonium sulphate cooperation precipitation	100	5.451	89.49	0.847	6.44	2.46	0.47
Acetone precipitation	10	4.847	79.58	0.461	10.51	4.01	0.83
Zinc sulphate precipitation	10	4.325	71.01	0.257	16.83	6.42	1.12

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 186 This purified SBP was introduced in various phenol concentrations viz. 250 mg/L, 500 mg/L,  
 187 750 mg/L and 1000 mg/L at various time intervals to evaluate the phenol degradation. The  
 188 per cent phenol degradation was determined based on residual phenol concentration. Figure  
 189 2 **on** effects of phenol concentration shows that 62.31% phenol degradation **was** obtained in  
 190 250 mg/L phenolic concentration at neutral pH after 8 h. As in microbial treatment, here also  
 191 **it was** observed that as phenol concentration increases the phenol degradation decreases.  
 192 Hence, only 21.82% phenol degradation **was** observed in 1000 mg/L phenolic concentration  
 193 at neutral pH after 8 h but this 21.82% phenol degradation is more as compared to microbial  
 194 treatment. Pradeep *et al.*, also gave a treatment of SBP on phenolic wastewater. They  
 195 obtained 72% phenol degradation of 100 mg/L phenolic solution. We had 62.31% phenol  
 196 degradation in 250 mg/L concentration, which was more.

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199  
200 **Fig. 2. Phenol degradation by enzymatic treatment**

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202 **3.3 Photocatalytic Treatment**

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204 **3.3.1 Effect of pH condition**

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206 Some properties of photocatalysts are highly pH dependent. Hence phenol degradation at  
207 different pH was carried out under UV light. In this treatment, TiO<sub>2</sub> nanoparticles were used  
208 as a photocatalyst. These nanoparticles were introduced at different pH (2-10) conditions to  
209 examine the phenol degradation. It is clearly seen that in figure 3, the basic conditions are  
210 unfavorable while acidic conditions are favorable for the photocatalytic degradation of  
211 phenol. In acidic medium, from pH 2 to pH 6 phenol degradation increases and after pH 6 it  
212 was decreaseds. The higher phenol degradation was observed with 63.08% at pH 6. The  
213 optimal pH condition was found to be acidic.

214  
215 Phenol has a pKa value of 9.95 and can be charged positively or negatively under the pH  
216 range studied; i.e., the attraction and interaction between both photocatalyst and phenol will  
217 be diverse with the solution pH. Moreover, as the pKa value of phenol is 9.95, it has negative  
218 charge above pH 9.95 ≈ 10 and referred as phenolate anions but the conversion of  
219 phenolate anions is commencing when solution pH in between 6 to 8 [26]. Conversely, in  
220 highly acidic condition phenol gets a positive charge while in weak acidic and neutral  
221 condition phenol molecules exist primarily in their non-ionic form. Additionally, the maximum  
222 OH<sup>•</sup> radicals are produced in the pH range of 6 to 7 [27], due to this reason rate of phenol  
223 degradation is higher in this pH range. These hydroxyl radicals, which are formed from some  
224 photocatalytic oxidative and reductive reactions. They have a capacity to directly break down  
225 of an aromatic ring of phenol molecule and transmute them into the final products which are  
226 CO<sub>2</sub> and H<sub>2</sub>O through various intermediates, because they are extremely strong, non-  
227 selective oxidants [28].

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229 **3.3.2 Effect of catalyst load**

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231 To examine the effect of TiO<sub>2</sub> nanocatalyst dosing on the phenol degradation, several  
232 experiments were carried out at catalyst loading from 1 to 4 g/L with 250 mg/L pollutant  
233 concentration. Figure 3 indicates that the increase in the amount of nanocatalyst loading  
234 also increases the rate of phenol degradation up to a particular catalyst dose of 3 g/L. This  
235 increased rate of degradation may be due to the higher surface area. Nevertheless, after 3  
236 g/L amount of catalyst loading the degradation rate starts declining. As the catalyst load  
237 increases, the experimental solution becomes turbid and resulting in UV rays getting  
238 scattered resulting-leading in-to a decrease in reaction rate [29]. The maximum phenol  
239 degradation at 3 g/L of catalysts dose was e-considered as an optimum condition for further  
240 study.

### 241 **3.3.3 Effect of H<sub>2</sub>O<sub>2</sub> and TiO<sub>2</sub> ratio**

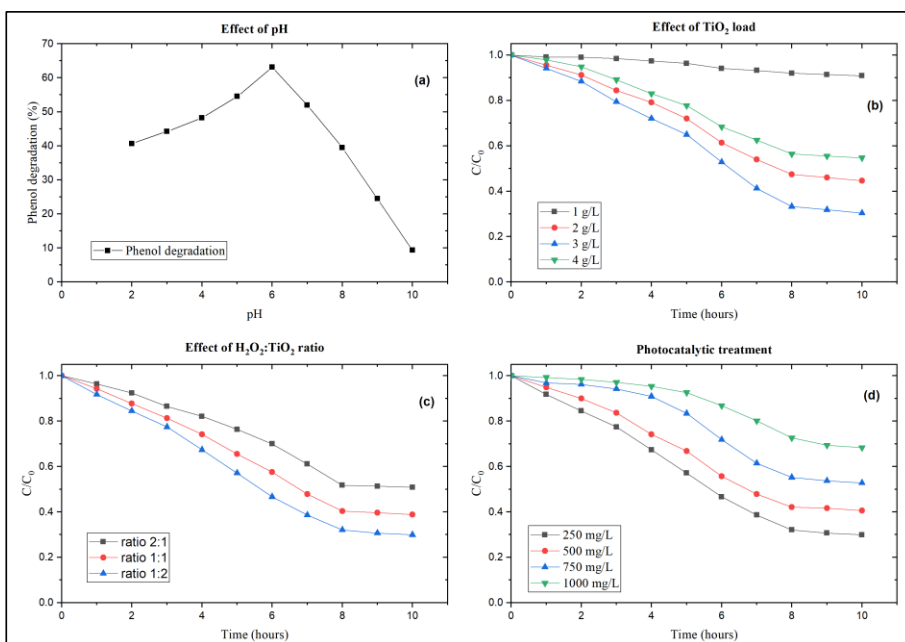
242 An oxidizing agent is another aspect of the photocatalytic oxidation process. Other  
243 experimental sets were performed for the study of the impact of various rates between H<sub>2</sub>O<sub>2</sub>  
244 and catalyst load as 2:1, 1:1 and 1:2. Figure 4 shows that a maximum phenol degradation  
245 was recorded at 1:2 ratio. It happens obviously because half the quantity of H<sub>2</sub>O<sub>2</sub> as on  
246 catalyst dose was enough for phenol degradation. The H<sub>2</sub>O<sub>2</sub> used only an oxidizing agent in  
247 a reaction medium. There is no use of a double quantity of H<sub>2</sub>O<sub>2</sub> in a-the reaction mixture.  
248 Because in an excess amount of H<sub>2</sub>O<sub>2</sub> reacts with those hydroxyl radicals which are  
249 responsible for degrading the pollutant molecule [30]. While the same quantities of H<sub>2</sub>O<sub>2</sub> and  
250 catalyst load, also not well for the degradation because there is no sufficient amount of  
251 catalyst in a-the mixture. This phenomenon also-was reported as earlier in 2001 by Ghaly *et*  
252 *al.*

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### 253 **3.3.4 Effect of phenol concentration**

254 TiO<sub>2</sub> nanoparticles applied in various phenol concentrations viz. 250 mg/L, 500 mg/L, 750  
255 mg/L and 1000 mg/L at various time intervals to evaluate the phenol degradation. The per  
256 cent phenol degradation was determined based on residual phenol concentration. As initial  
257 phenol concentration increases, the rate of phenol degradation decreases from 250 mg/L to  
258 1000 mg/L. This happens due to the competitive adsorption on the active sites of  
259 photocatalyst between the hydroxide radicals and phenol molecules [31]. Figure 4 on effect  
260 of phenol concentration shows that 68.39% phenol degradation obtained in 250 mg/L  
261 phenolic concentration at neutral pH after 8 h. As in microbial treatment, here also seen that  
262 the phenolic concentration increases the phenol degradation decreases. Hence, only 28.46  
263 % phenol degradation observed in 1000 mg/L phenolic concentration at neutral pH after 8 h,  
264 but this 28.46% phenol degradation is more than in microbial treatment. Pradeep et al. also  
265 gave a treatment of SBP on phenolic wastewater. They obtained 72% phenol degradation of  
266 100 mg/L phenolic solution. We had 68.39% phenol degradation in 250 mg/L concentration,  
267 which was more.

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275 **Fig. 3. (a) Phenol degradation at various pH conditions, (b) Effect of TiO<sub>2</sub>**  
276 **nanoparticles loading on phenol degradation, (c) Effect of H<sub>2</sub>O<sub>2</sub>:TiO<sub>2</sub> nanoparticle ratio**  
277 **on phenol degradation and (d) Effect of different phenolic concentration on phenol**  
278 **degradation under UV light**

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280 **3.3.5 Degradation rate kinetics**

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282 The kinetic study of photodegradation of phenol was investigated for UV/H<sub>2</sub>O<sub>2</sub>/TiO<sub>2</sub> system.  
283 A model with a higher value of correlation coefficient (R<sup>2</sup>) considered as more applicable.  
284 The equation for first and second order kinetics is shown below.

285 First order reaction kinetics:  $\log(qe - qt) = \log(qe) - \left(\frac{K_f}{2.303}\right)t$  (2)

286 Second order reaction kinetics:  $\left(\frac{t}{qt}\right) = \left(\frac{1}{K_s qe^2}\right) + \left(\frac{1}{qe}\right)t$  (3)

287 Where q<sub>e</sub> and q<sub>t</sub> are the amounts of phenol degradation (mg g<sup>-1</sup>) at equilibrium time and at  
288 time t (min), respectively. K<sub>f</sub> is the rate constant of first-order reaction (min<sup>-1</sup>) which can be  
289 obtained from the slope of log (qe-qt) versus time plot. Also, a rate constant of pseudo-  
290 second-order K<sub>s</sub> reaction (g mg<sup>-1</sup> min) can be obtained from t/qt versus t plot. For the phenol,  
291 first-order reaction kinetic was fitted than second-order reaction kinetics first order having a  
292 maximum value of R<sup>2</sup>. Besides the apparent first-order rate constants decreased with the  
293 increase of initial phenol concentrations [32]. Hence, kinetic constant based on phenol  
294 degradation by UV calculated for a first-order reaction. Table no. 2 shows a description of  
295 first-order reaction kinetics.

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297 **Table 2. Description of first-order reaction kinetics**



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Substrate	Concentration (mg/L)	K (min <sup>-1</sup> )	R <sup>2</sup>
Phenol	250	0.0953	0.9838
	500	0.0555	0.9793
	750	0.0088	0.8960
	1000	0.0067	0.8546

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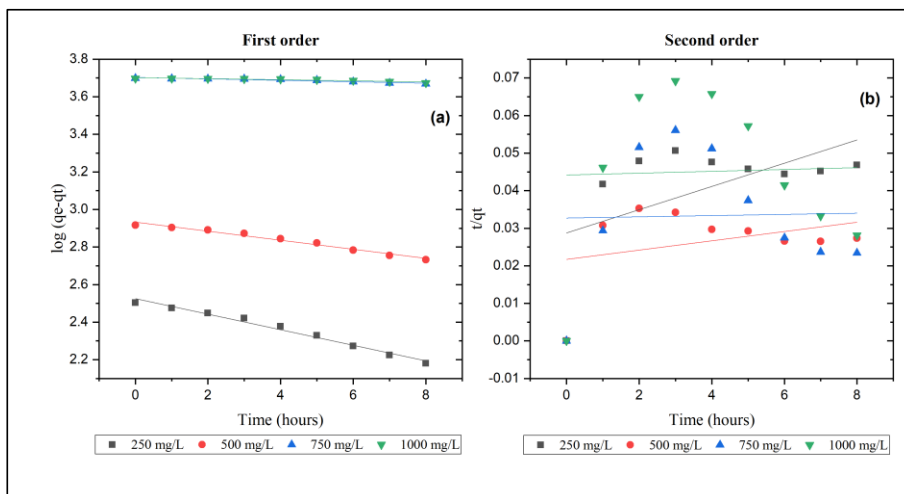


Fig. 4. Phenol degradation corresponds to the (a) first-order and (b) second-order model for 250, mg/L, 500 mg/L, 750 mg/L and 1000 mg/L

#### 4. CONCLUSION

This study adopted three methodologies ~~such as which were~~ microbial, enzymatic and photocatalytic treatments of phenol for the degradation. Microbial treatment gives 60.07%, enzymatic treatment gives 62.31%, and photocatalytic treatment gives 68.39% phenol degradation in 250 g/L phenolic concentration. All treatments gave approximately the same phenol degradation, but each treatment has some advantages as well as some disadvantages. About 60.07% phenol degradation achieved under 96 h in microbial treatment whereas 62.31% and 68.39% phenol degradation takes place under 8 h in enzymatic and photocatalytic treatment. Based on the time parameter, microbial treatment is a very time-consuming method for phenol degradation while ~~remaining both the other~~ methods are less time-consuming.

In enzymatic treatment, additional one-step is required for phenol degradation. That step was enzyme purification. Enzyme purification method was adopted in this study, and that the purified enzyme used as a catalyst. An enzymatic treatment did not show significant phenol degradation even after purified enzyme was introduced in ~~a the~~ reaction mixture. In phenol degradation follow another one-step and degrade the phenol which is not much more. Therefore, this enzymatic treatment is not a feasible method for phenol degradation.

A remaining method is a photocatalytic degradation. It requires less time, no need for extra steps. The maximum phenol degradation achieved in this photocatalytic method, i.e. was

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327 | 68.39%. ~~A~~The whole photocatalytic ~~study~~ degradation was performed under acidic  
328 | condition, this is one thing which is noticeable. However, there is no need of extra handling  
329 | of that acidic medium. Overall, from the comparative study of all the three methods reported  
330 | in this study, the photocatalytic process is useful efficient for phenol degradation than others.

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### COMPETING INTERESTS

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Authors have declared that no competing interests exist.

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