

Hydrogeological characteristic and the vulnerability degree of the aquifers from Municipality of Abaetetuba, Pará - Brazil

ABSTRACT

This study evaluated the quality of groundwater in the municipality of Abaetetuba (PA, Brazil 1°43'46" S e 48°52'27" W) based on the hydrogeological characterization and degree of vulnerability of the aquifer system. The municipality of Abaetetuba is practically all supplied by groundwater both by deep tubular wells and shallow pit wells (Amazonian wells), which present potential risk of contamination. Water and soils samples from 20 wells sampled between 2012 and 2016 were used. Physicochemical and microbiological analyzes served as a data base for mapping (GIS). Three hydrogeological domains were identified within the study area: Barreiras Group (predominant), Post-Barrier Sediments and Recent Sediments. Almost all of the samples presented microbiological levels above the MPV defined by the Brazilian legislation for water intended for human consumption. The DRASTIC and GOD vulnerability indexes presented values between 75 and 119 and 0.15 and 0.32 respectively, suggesting areas of 'low' to 'moderately-high' vulnerability. Local sources of pollution by coliforms have been identified due to lack of basic sanitation. Evidence of diffuse sources derived from vehicle washing and lubrication also have been identified. Although the Barreiras Group had deep semi-confined aquifers, a 'state of alert' was suggested for the areas that indicated the presence of fecal coliforms and with a high population density. Isovalues and vulnerability maps suggest areas that require further monitoring. A positive correlation between the DRASTIC Index and TDS in well water was established. The intense exploitation of groundwater especially in areas of great population density may be causing contamination of aquifers.

Keywords: groundwater pollution; DRASTIC; GOD; GIS; aquifer vulnerability; Amazon

1. INTRODUCTION

Approximately one-fifth of the world's total fresh water lies in the saturated zone of the groundwater environments [1], and hence needs to be protected from any contamination. With the exponential increase in demand for potable water, the number of wells drilled for supply each year increases further. Historical data show that in the last 30 years more than 300 million wells have been drilled in the world [2]. The US drills an average of 400,000 wells per year, with more than 200 million m³ used by the Midwestern states of the country, in the arid and semi-arid zone of Nebraska to Texas [3]. Mexico City, one of the most populous cities in the world, is practically supplied by wells, and 75% of the European Community has public systems supplied by groundwater, reaching 90% in countries such as Germany, Belgium, Sweden and Denmark [4,5]. In Brazil, the lack of proper registration and control makes it difficult to establish a more accurate estimate of water demand. Despite this, it is estimated that more than 60% of the Brazilian population is being supplied by underground water, 45% of which is through tubular wells [6]. Cities such as Ribeirão Preto (SP) and Mossoró (RN), and state capitals such as Maceió (AL), Manaus (AM), Natal (RN) and Recife (PE) are practically supplied only by tubular wells [6,7]. In addition to serving the population directly, these resources are used in industry, agriculture (irrigation) and leisure. In the most

28 populous region of Brazil (Southeast) between 75 and 90% of the population of the cities is
29 supplied by wells [6,8].

30 Groundwater found in aquifer systems is stored water that accumulates over thousands of
31 years, which under natural conditions are in a balanced state, controlled by recharge-
32 discharge mechanisms and by the potentiometric load difference between the system fluvio-
33 lacustre and underground. These waters are influenced directly by the climate (precipitation)
34 and by the degree of permeability of the soils. Its movement underground is very slow,
35 implying a high residence time. Contrary to the general idea, most groundwater is found in
36 deep aquifers, of the non-draining confined type, whose upper and lower boundary layers
37 are impermeable or semi-impermeable, limiting their use by deep wells.

38 The increasing global demand for groundwater combined with the urbanization process,
39 which increases the area of paving and macrodrainage of public roads, reduces the flow of
40 recharge, putting the aquifer reserves at risk. Reducing the recharge flow through the paving
41 of large cities may cause lowering of the water table and the saturation zone of the aquifers
42 themselves, and it is necessary to drill deeper wells to obtain a satisfactory flow. The quality
43 of the groundwater is also being harmed by the infiltration of pollutants and contaminants
44 through the soil (authors' note). The use of groundwater is becoming increasingly
45 problematic due to lack of planning and lack of basic sanitation systems, especially in third
46 world countries. Another problem to be solved is the high cost of the treatment system,
47 necessary to meet drinking standards. In Brazil, the drinking standards are very discerning
48 and defined by the Ministry of Health. Despite this, due to the high demand and the high cost
49 of treatment, most of the tubular well water is distributed to the population without previous
50 treatment, only with chlorination simple.

51 The Amazon region has an immense hydrological reserve resulting from the largest
52 hydrographic basin and the largest underground aquifer in the world. However, the lack of
53 basic sanitation services, together with the infiltration of pollutants and contaminants from
54 dumps, cemeteries, gas stations, etc., undermine the use of this resource. The municipality
55 of Abaetetuba in the State of Pará is practically all supplied by groundwater both by deep
56 tubular wells and shallow pit wells (Amazonian wells), which are generally constructed
57 without inspection, norms and environmental license, and presenting a potential risk of
58 contamination. Studies developed in the municipality from Georadar (GPR) indicate signs of
59 contamination of soils by hydrocarbons, which increase the risk of contamination of
60 wastewater to the population [9]. The objective of this research was to evaluate the quality of
61 groundwater in the Municipality of Abaetetuba (PA, Brazil) based on the hydrogeological
62 characterization and degree of vulnerability of the aquifer system. A technical-qualitative
63 analysis of 20 tubular wells sampled between 2012 and 2016 was performed. Physical-
64 chemical and microbiological analyzes of groundwater (alkalinity, acidity, total hardness,
65 electrical conductivity, ionic concentration [Na^+ , K^+ , Ca^{2+} , Mg^{2+} , Cl^- , Cl_{dpt}^- , SO_4^{2-} , HCO_3^- and
66 CO_3^{2-}], total coliforms and thermotolerant coliforms) served as a data base for making
67 thematic maps, with the help of Geographic Information System tools. The results of the
68 mapping were correlated with the use and occupation, soil type and socioeconomic impact
69 for the region.

70 2. STUDY AREA

71 2.1. Physical aspects

72 The municipality of Abaetetuba ($1^\circ 43' 46''$ S and $48^\circ 52' 27''$ W, Fig. 1) belongs to the
73 northeast mesoregion of the Pará State, 22 meters above sea level. Considered a pole-
74 town, Abaetetuba is located 51 km S-W of Belém, and makes borders to the north with the

city of Barcarena, and to the south with the cities of Moju and Igarapé-Miri. The municipality has a territorial area of 1610.6 km², being 57% of the total occupied by rural areas. According to Köppen's classification, the climate of the region is hot and humid, with high temperatures (27 °C average) and high precipitation (2000 mm/year) [10,11]. The soils of the region are influenced by temperature and precipitation, causing leaching. There is a predominance of clay and sandy-clay soils, with low organic matter content and high aluminum concentration, which intensifies cation exchange processes [12]. According to the classification of the soils [13], predominantly Latosols Yellow Dystrophic soils associated with Hydromorphic Podzol and Concrete Lateritic soil are found. Gley soils and Eutrophic and Dystrophic Alluvial soils, especially in recent soils on beaches and islands, can also be observed. The relief of the region is flat and low with several coastal plains. The platform areas oscillate between 15 and 30 meters (average 20 m), and there are also lowland areas (<4 m), where occurs flooding. The primitive vegetation typical of the Amazonian practically no longer exists. The hydrographic network of the municipality is quite vast, sinuous and of strong asymmetry with several navigable stretches. The main rivers are Pará, Tocantins, Abaeté, Guajará de Beja, Arapiranga and Arienga [10,13].

2.2. Socioeconomic Aspects

The resident population of Abaetetuba in 2010 was 141,100 inhabitants with a population density of 87.6 inhabitants per km². By 2017 it was estimated a population of 153,380 inhabitants [6,10,11], which represents an increase rate of 8.7% in 7 years. The age pyramid is classic of third world countries, with a large base composed of young people between 15 and 25 years old, and a narrow top from 70 years old. The main economic activities in the region are commerce and services of the most diverse activities. Industrial activity has a small share in the local economy, but has been showing great growth in recent years, especially in the food and agro-forestry products sectors [11,14]. The schooling rate for children between 6 and 14 years old in 2010 was 97.7 and the average infant mortality rate was 9.25 deaths per 1000 live births [11]. The Municipal Human Development Index (MHDI) has grown in recent years, from 0.628 in 2010 to 0.751 in 2013 [6,15]. Despite this, socioeconomic development is still modest, and considering the aspects of basic sanitation, such as the supply of treated water, sanitary sewage, and rainwater and solid waste management, the municipality of Abaetetuba presents conditions that are still very precarious. Most of the population still does not have regular water supply. The poorer population excavates their own wells, generally not following building and well preservation standards, making the wells vulnerable to contamination.

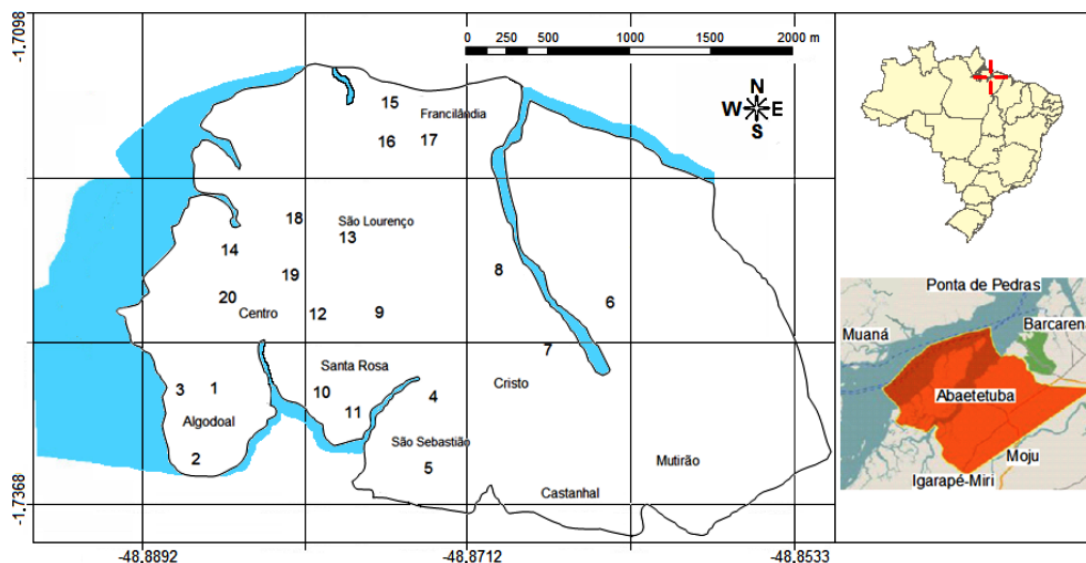


Fig. 1. Location of the municipality of Abaetetuba (Pará, Brazil) with indication of monitored tubular wells. (Source: CPRM/SIAGAS 2018 adapted in ArcGis © 9.3).

3. MATERIALS AND METHODS

A technical-qualitative analysis of 20 tubular wells, sampled between 2012 and 2016, was performed using the physical-chemical (alkalinity, acidity, total hardness, electrical conductivity and ionic concentration) and microbiological (total and fecal coliforms) data determined by [16]. Technical information on the wells in the study area was obtained through contact with public and private sector companies, accessing technical reports from drilling and/or groundwater management companies, basin water user registry, and institutions of research of the respective area. Data provided by the SiAGAS/CPRM on registered wells (SIAGAS, Fig. 2) [17] were consulted to aid in the hydrogeological characterization and geometric profile of the wells monitored. For the general configuration of the results, based on the physicochemical and microbiological patterns, the hydrogeological characteristic and the degree of vulnerability, the studies were mainly concentrated in the free aquifers of alluvial sediments of the Quaternary and semi-confined aquifers of Tertiary sediments of the Barreiras Group. The ionic balance $[Na^+, K^+, Ca^{2+}, Mg^{2+}, Cl^-, SO_4^{2-}, HCO_3^- \text{ and } CO_3^{2-}]$ expressed in meq/L and the ionic classification of groundwater were determined from the hydrochemical diagram of Piper [18] and Stiff diagram [19]. The data was processed using ExPiper[®], Microcal Origin[®] 9.0 and AquaChem 3.7 software's. The hydrogeological classification included the analysis of the hydraulic conductivity, grain size, flow direction and relative permeability. Sampling quality control for ionic analysis was performed based on the ion balance (Ib%) calculation defined by [20] in equation 1.

The analyses were processed in the chemistry laboratories of UFPA/ICEN. In the sampled soils, gravimetric humidity (Gh %) and volumetric humidity (Vh %) were determined by difference of mass before and after the oven drying at 105 °C (equations 2 and 3), and the density (kg/dm^3) were determined by the ratio gravimetric and volumetric humidity (equation 4). The results were compared with the calculation of the mass/volume ratio of the sampled soil using metallic cylindrical of $9.812 \times 10^{-2} dm^3$ volume (equation 5). The total porosity of the soil was estimated by the volume ratio of the solid and volume of saturation (equation 6, Table 1). The granulometric analysis was determined by the fractionated sieving method,

141 using the TYLER series sieves with mesh openings between 2.0 and 0.032 mm. The
142 percentage of silt and clay fractions ($\phi < 0.063$ mm) were determined by wet sieving with
143 sodium hexametaphosphate $[\text{NaPO}_3]_n \cdot n\text{Na}_2\text{O}$ as dispersing agent. The saturation state of
144 the soils was estimated by the application of a well-known volume of distilled water in a
145 regular metal cylinder filled with soil. The analyzes followed recommendations and protocols
146 from [21,22]. The hydrogeological classification was elaborated from the protocols described
147 by [23], which takes into account the conversion of the geological units into hydrolytic units,
148 differentiating the units in continuous or discontinuous, according to their geometric
149 characteristics and forms of occurrence. Another possibility that the protocol considers is to
150 identify the lithological types according to the dominant flow characteristics. The
151 nomenclatures described to Belém and Ananindeua areas [24] and to the Legal Amazon [25]
152 were applied. The geological uniformity of the mesoregion was considered, as a
153 consequence of the territorial proximity and similar geometric configurations between the
154 aquifers. The area between Belém Metropolitan Region (Belém-Ananindeua) and the
155 municipalities of Bacarena and Abaetetuba (axis NE-SW) was defined as mesoregion.

156 **Table 1: Summarized analysis of the calculations applied to the soils sampled in the**
157 **wells.**

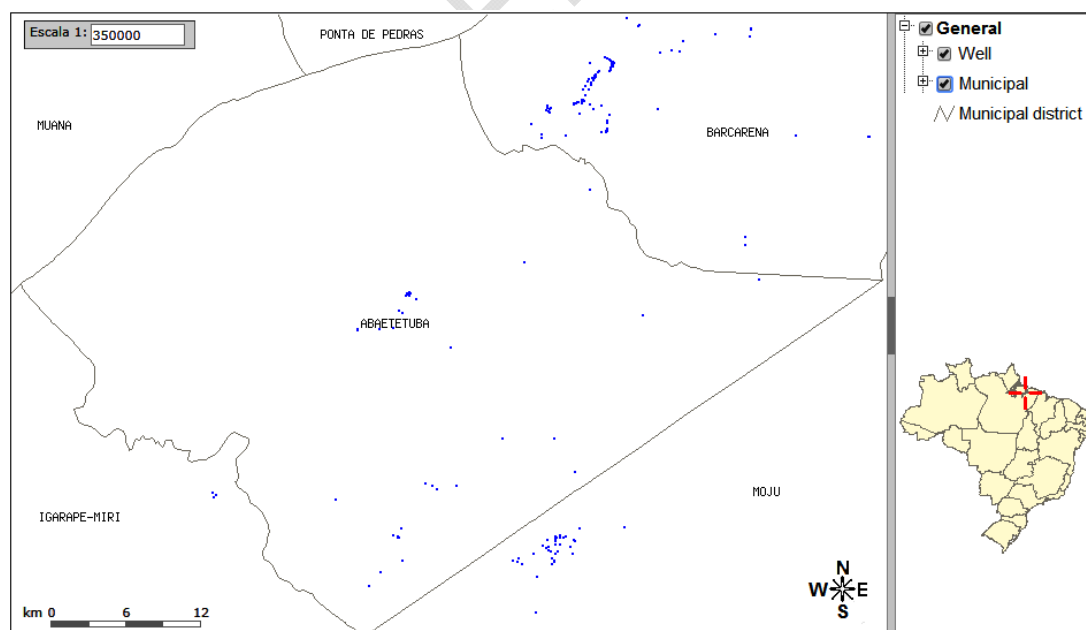
Analysis	Equation (eq.)	Description
Ion balance	$eq.1: Ib(\%) = \left \frac{\sum anions - \sum cations}{\frac{1}{2}(\sum anions + \sum cations)} \right \times 100$	
Gravimetric humidity	$eq.2: Gh(\%) = \left(\frac{a-b}{b} \right)$	$a =$ wet mass; $b =$ dry mass
Volumetric humidity	$eq.3: Vh(\%) = \left(\frac{a-b}{c} \right)$	$c =$ volume of the sample
Density	$eq.4: d(kg/dm^3) = \frac{Vh}{Gh}$	
Mass/volume ratio	$eq.5: d(kg/dm^3) = \frac{soil\ mass}{cylinder\ volume}$	volume = $9.812 \times 10^{-2} dm^3$
Total porosity	$eq.6: P(\%) = 1 - \frac{solid\ volume}{saturation\ volume} = 1 - \frac{d}{d_{mineral}}$	$d_{mineral} = 2.66 g/cm$
DRASTIC Index*	$eq.7: DI = D_{RW} + R_{RW} + A_{RW} + S_{RW} + T_{RW} + I_{RW} + C_{RW}$	
GOD Index*	$eq.8: GI = G_w \times O_w \times D_w$	

158 *where the capital letter indicates the corresponding parameter; the subscript 'R' and 'W' refer to the
159 variable rating and weight factor, respectively to [27] and [29].

160 Several groundwater vulnerability assessment methods have been developed. The most
161 part of that methods divide groundwater vulnerability assessment methods into three
162 categories such as overlay and index methods, methods employing process-based
163 simulation models and statistical methods [26]. The natural vulnerability of the groundwater
164 studied was evaluated using the DRASTIC method [27], which takes into account the sum of
165 the seven hydrogeologic factors of the region: depth to groundwater table (D m); aquifer
166 recharge (R mm/year); aquifer media (A); soil media (S); topography (T %); impact of
167 vadose zone (I), and hydraulic conductivity (C m/day; equation 7, Table 1), which are a
168 combination of geologic, hydrogeologic, geomorphologic, and meteorological characteristics
169 of an aquifer. In this study, depth to groundwater table (D) was obtained by subtracting the
170 ground surface elevation from the average groundwater level of observation wells. The

171 average annual net recharge of the aquifers (R) at Abaetetuba was estimated based upon
 172 Water Table Fluctuation method (WTF), where the average net recharge value varies from
 173 40 to 60 mm/year. Aquifer media (A) was obtained using the available information on
 174 geological cross sections, geological survey and drilled well logs data. Soil media (S) was
 175 obtained based on available soil maps and grain size analysis of borehole samples of the
 176 region. Slope information of the topography (T) was extracted from the CPRM/SIAGAS [17].
 177 The depth from ground surface to groundwater table in the study area is variable (25 to 90
 178 m) with depth aquifers belonging mainly to Barreiras Group. Thus, the thickness of soil zone
 179 and thickness of remaining part of vadose zone were considered to estimate the impact of
 180 vadose zone (I). The values of hydraulic conductivity (C) were obtained from field pumping
 181 tests data and sieve analysis. The DRASTIC model developed by [27] is the most usual
 182 vulnerability mapping method, used as important instrument for groundwater planning and
 183 decision making. The final DRASTIC indexes can range from 26 (zero vulnerability) to 226
 184 (very high vulnerability) according to [28]. To determine the vulnerability index by
 185 contamination, applied to isovalues maps, the GOD model proposed by [29] was used.
 186 Foster and Hirata [29] established this index from the product of three parameters (equation
 187 8, Table 1): groundwater occurrence (G), which represents the type of occurrence of
 188 groundwater, with indices ranging from 0.0 (confined aquifers) to 1.0 (free aquifers); overall
 189 of litology of aquiperm (O), determined by geological mapping of the unsaturated zone and
 190 the lithological profile of the well, with values ranged between 0.3 and 1.0; and groundwater
 191 depth (D), referring to the depth of the static level, ranging from 0.3 to 0.9. The result is an
 192 index capable of identifying the vulnerability levels of the aquifers associated to the installed
 193 wells, and expressing their degree of natural resistance to the penetration of contaminants.
 194 To confirm the degree of vulnerability, the water quality standards for human consumption
 195 and their drinking water standard were applied, based on the Permissible Maximum Value
 196 (PMV) described in Ordinance N° 2914/2011 of the Ministry of Health [30]. The results were
 197 presented on maps of isovalues with interpolation using Surfer® Golden Software, 9.11
 198 (2010) and ArcGis® 9.3 (2008) ESRI - USA.

199



200 **Fig. 2. Geopolitical limits of the municipality of Abaetetuba (Pará, Brazil) with the wells**
 201 **registered in the CPRM - Geological Service of Brazil.** (Source: CPRM/SIAGAS 2018
 202 adapted in ArcGis © 9.3).

203 4. RESULTS AND DISCUSSION

204 4.1. Geological and hydrogeological characterization

205 The investigation of the geological and hydrogeological aspects is fundamental for the study
206 of the ionic composition of the waters, as well as to establish levels of vulnerability for the
207 aquifers. In the Pará State, the Pirabas Formation, Barreiras Group, Post-Barrier Sediments
208 and Recent Sediments (Quaternary) stratigraphic units were identified. These units are
209 distributed in a discontinuous way, being possible to observe punctual and irregular outcrops
210 in the macroregion. The Pirabas Formation (Miocene - Oligocene) is predominantly found in
211 the NE direction of the Pará, and in parts of the cities of Belém and Ananindeua. Because it
212 presents a transition to marine sediments, it is also found in the direction of the Marajó
213 Island and in the coastal strips of the states of Maranhão and Piauí. The Barreiras, Post-
214 Barrier Sediments and Quaternary Coverage geological units were identified in the
215 mesoregion where the municipality of Abaetetuba is located. The Barreiras Group is used as
216 a generic term for continental sediments. It is a designation applied to indicate non-
217 fossiliferous clastic sediments friable and of intense colors that occurs almost uninterruptedly
218 along the Brazilian coast, North and Northeast regions and in deposits of the Amazon valley
219 [31]. It corresponds to ferruginous sandstones, siltstones, argillites and colored clays and
220 unconsolidated clay-sandy and sandy-clay sediments. The sediments of the Barreiras Group
221 usually have fine to medium granulometry, and can accumulate forming both cliffs in the
222 coastal zone and sandy-clay banks on the banks of the Amazon Basin. Those sediments
223 may also eventually be found as coarse sediments to conglomerates. According to [13,14],
224 in Pará the Barreiras Group is adjusted over the Miocene to Mio-Pleistocene rocks. Post-
225 Barrier Sediments belong to the Pleistocene and are associated with yellow sediments,
226 unconsolidated, superimposed on the layers of the Barreiras Group. It consists of sandy-clay
227 sediments, of medium to fine granulometry (from quartz to clay) with ferruginous
228 concretions. Post-Barrier Sediments occurs primarily in the areas of floodplain, igarapés and
229 islands especially in the direction NE, already in the limits with the municipality of Bacarena.
230 Recent Holocene sediments are composed of unconsolidated sediments containing clays of
231 reddish color and sandy of gray to brown coloration, with very variable granulation and
232 presence of vegetal remains.

233 The alluvial processes derive from erosion - transport - sedimentation, acting on the flow of
234 debris typical of alluvial fans or fluvial channels [31 and Note of the authors]. Of an
235 essentially clastic nature, this stratigraphic unit is mainly found in the beaches of rivers and
236 streams (flood plains) in the western boundary of Abaetetuba and in the Marajó and Guajará
237 bays, in the direction NE of the municipality, associated to Coastal Deposits (Fig. 3). It is
238 estimated that in the urban zone these sediments can present thicknesses of the order of
239 maximum 10 m. The geological and geomorphological features exert a great influence on
240 the hydrogeological conditions of the aquifers.

241 The city of Abaetetuba is predominantly inserted in the Pará River basin, presenting in the
242 subsoil a significant reserve of water. Considering the protocols established by [17],
243 consultations to geological and hydrogeological maps and from observations in the local
244 drilling were identified in the mesoregion the hydrogeological domains Barreiras, Post-
245 Barreiras and Aluviões (Table 2). These results are confirmed by the studies of [25] to
246 elaborate the hydrogeological maps of the Legal Amazon. The Barreiras domain is located in
247 most of the limits of the municipality of Abaetetuba, except for the lowland areas (Fig. 3),
248 where alluvial sediments predominate. It is the main aquifer captured by the wells of the
249 region. They are aquifers generally confined to semi-confined, depending on the degree of
250 discontinuity of the lower and upper layers. It presents a matrix of heterogeneous
251 granulometry and variable thickness, and its recharge occurs by contribution of the

overlapping layers, or through precipitation in the outcrop areas. The Barreiras aquifer presents depth of occurrence between 25 and 90 m, and flow from 10 to 70 m³/h. The Post-Barreiras hydrogeological domain is characteristic of free to semi-confined aquifers (free with coverage) and with variable depth, but generally less than 25 m and flow less than 5 m³/h. These characteristics make the wells vulnerable to sources of contamination, which can reach the water table. The Post-Barreiras domain consists of alluvial sands, as well as fine to medium sandy and clay-sandy materials. The aquifers of Recent Sediments can be divided into free or free with coverage, whose recharge occurs directly through rainfall. Its depth of occurrence is less than 10 m, and shallow wells excavated in the lowland areas belong to this hydrogeological domain. The discharge occurs through rivers, springs and wells, with an average flow 10 m³/h. The alluvium is a permoporous domain, with good water storage capacity, but not significant in the region of Abaetetuba. In these cases, a vertical sequence of coarse sandy and pebbles at the base up to silts and clays at the top was observed, with a thickness of 10 m. It was also identified the Marajó Domain [17] (Fig. 3), towards the Atlantic Ocean, on the N-NE axis of the continental area. These are miocene and postmyocernic sedimentites with abundant plant remains, deposited by the influx currents from the Marajó basin [31 and Note by the authors]. In periods of high tide, the strong inflow transports the sedimentitos to the channels of the water network of the mesoregion Belém - Abaetetuba.

Table 2. Hydrogeological domain of the aquifers found in the municipality of Abaetetuba.

Domain	Epoch	Type	Location
Barreiras	Mio-Pleistocene	Confined to semi-confined	most part of municipality
Post-Barreiras	Pleistocene	Free with coverage	Várzea, igarapés, islands
Aluviões	Holocene	Free	Floodplain

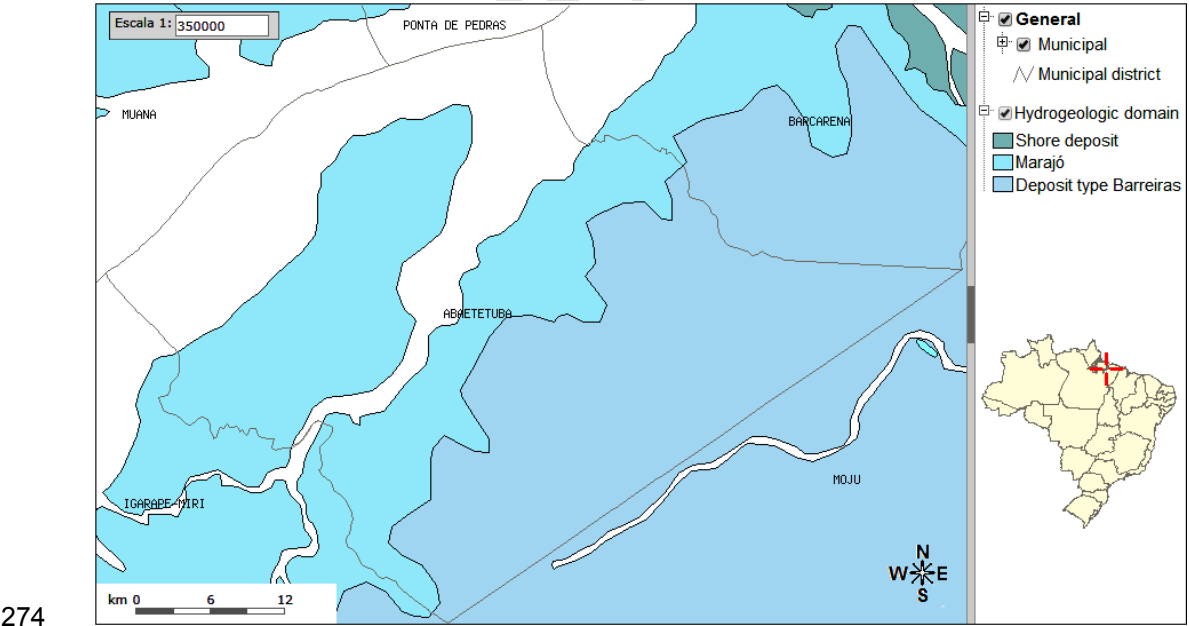
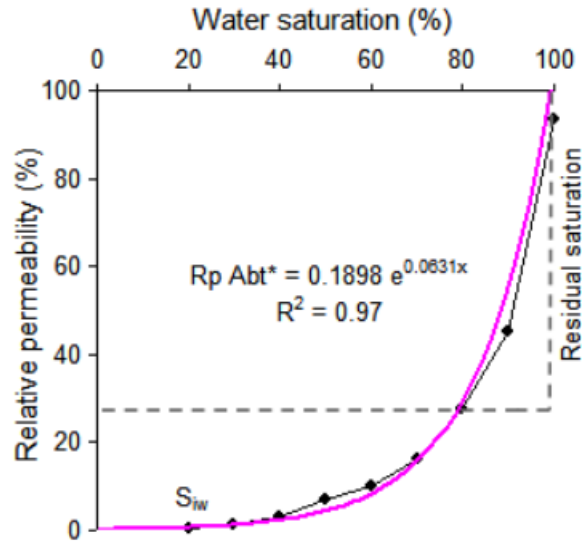


Fig. 3. Hydrogeological domains observed in the limits of the municipality of Abaetetuba (Pará, Brazil). (Source: CPRM / SIAGAS 2018 adapted in ArcGis © 9.3).

277 The sampled sites comprise the districts of Francilândia, São Lourenço, Centro, Algodão,
278 Santa Rosa, São Sebastião and Cristo (Fig. 1). The majority of the wells monitored are
279 located in a semi-confined aquifer, with the exception of sites 15 and 16, which indicate the
280 presence of free to free with coverage. The results of the granulometric analysis suggest
281 predominance of a surface layer of sandy-clay soil with similarity between sites 4 to 9 (I-S
282 group), 13 to 16 (group II-S) and 18 to 20 (group III-S); of yellow soil for sites 10 to 12 (group
283 IV-S), and sandy soil at sites 1 to 3 and 17 (group V-S). The results for the bottom sediments
284 suggest a predominance of clay soils with similarity between the sites 1 to 12, 15 and 16
285 (group I-F); and sandy coarse soils at sites 13 and 14 and 17 to 20 (group II-F). The physical
286 environment (soil texture) indicates the greater or lesser potential of infiltration capacity of
287 organic and inorganic contaminants, including microorganisms, in the sedimentary column.
288 In this case, the highest resistance to percolation of contaminants in the first layers of soil
289 was observed in the Santa Rosa neighborhood (sites 10 to 12, Fig. 1).

290 From the grain size and degree of soil saturation results, the relative permeability (R_p) of the
291 soils in the wells monitored was estimated as a function of residual saturation. The
292 permeability coefficient of a porous medium is dependent on the type of liquid present in the
293 system (usually water). There is also the possibility of different non-miscible liquids (e.g.
294 water and oil). Also interfere in the degree of permeability the type of texture of the soil. For
295 the soils of the study area, the calculations suggest that 100% water saturation is obtained
296 with 93.4% R_p ; while at 50% saturation only 6.9% of soil R_p is needed (Fig. 4). The
297 coefficient of R_p decreases exponentially until it reaches the point S_{iw} (irreducible saturation
298 point of water), when R_p tends to zero and water becomes effectively immobilized.
299 Knowledge of the degree of relative permeability of soils helps in understanding the
300 mechanisms of pollutant and contaminants transfer in the sedimentary column. The
301 dynamics of the contaminants in the soil is explained through mass transfer mechanisms by
302 advection, dispersion and/or attenuation [32]. The advection consists of the mechanism
303 where the contaminants follow the vectorial flow, presenting a direct relation with the velocity
304 of percolation in the soil. The dispersion consists of the mechanism responsible by the
305 reduction of the pollutants concentration in the percolation fluid, be for hydrodynamic
306 dispersion or molecular diffusion. The attenuation is the reduction of pollutants transported
307 by the advection or dilution of these from physical-chemical and biological reactions [32 and
308 Note by the authors]. In this case, the concentration of oxygen in the interstitial waters is the
309 regulating factor of the oxidation reactions, especially of the organic compounds. Of the
310 three mass transfer mechanisms, the attenuation is the one that requires more attention,
311 because it can be temporary. This occurs when the pollutants are not reduced from the
312 oxireduction reactions, being simply adsorbed to the soil particles. This is an important
313 aspect in the case of metallic elements, whose availability in the water column depends
314 directly on pH and alkalinity [12,33]. Considering an efficient buffer system, without large
315 variations of the alkalinity, the metal ions are adsorbed to the clay particles [12] and remain
316 suspended in the total solids. This can compromise the quality of the well water.



317

318 **Fig. 4. Water permeability curve for the 20 monitored wells (mean - black line) and**
 319 **trend curve (pink line). S_{iw} = irreducible saturation point of water; Abt = Abaetetuba.**

320 4.2. Physical-chemical and microbiological characterization

321 The waters of the analyzed wells are mineralized with values of electrical conductivity
 322 varying from 35.8 to 295.3 (average 128.1 ± 73.3 $\mu\text{S}/\text{cm}$). The hydrogeochemical results
 323 presented in the Piper Diagram (Fig. 5A) indicated the predominance of calcium chlorinated
 324 groundwater with concentrations ranging from 3.5 to 102.1 (43.6 ± 39.4 mgCa/L) and from 1.8
 325 to 68.9 (19.0 ± 17.6 mgCl/L), and calcium bicarbonate waters with concentrations of $\Sigma(\text{HCO}_3^-$
 326 and $\text{CO}_3^{2-})$ varying between 2.6 and 74.8 (mean 30.7 ± 28.1 mg/L). The oscillation found in
 327 the ionic concentration is due to small variations in ionic levels in the soils of the area. For
 328 the cations, the Ca^{2+} and Mg^{2+} contents did not vary with the depth of the well, while the Na^+
 329 decreased gradually with the increase of the depth. The anion contents remained stable in
 330 the water column, except for HCO_3^- that gradually decreased with increasing depth. The
 331 levels of HCO_3^- , CO_3^{2-} and Ca^{2+} observed in the waters demonstrate natural processes of
 332 interaction between water and rock in the carbonate domain. Cl^- observed in the wells may
 333 be associated with climatic factors such as precipitation followed by percolation of the
 334 adjacent soil. Since chlorine is not very abundant in the rocks that make up the geology of
 335 the Metropolitan Region of Belém, its occurrence in groundwater must be associated with
 336 the rainwater and the influence of the sea through the proximity of the brackish waters of the
 337 Marajó Bay, to the West of the area studied [24]. The Cl^- ion is present in all natural waters,
 338 usually from the leaching of ferromagnesian minerals from igneous rocks and evaporitic
 339 rocks. Groundwater has Cl^- normally less than 100 mg/L . The various natural interactions
 340 experienced by groundwater generally do not influence the increase of this factor. Therefore,
 341 the increase in chloride concentrations can mean anthropic activities, especially industrial
 342 waste and sanitary sewage.

343 A spatial similarity analysis (Cluster with WPGA) of the ionic balance confirmed the presence
 344 of four classes of wells: group I-S formed by sites {1,3,9,12 and 13} with more chlorinated
 345 and less calcic water; group II-S {2,11,14,16,18 and 19} with more calcic and less
 346 chlorinated waters; and groups III-S {4,7,8 and 10} and IV-S {5,6,15,17 and 20} with
 347 intermediate concentrations for Cl^- and Ca^{2+} ions. The ionic concentrations in meq/L
 348 represented on lines in the Stiff Diagram confirm the results of the ion balance, with an

349 elongated hexagon with vertices in the calcium and chloride ions (Fig. 5B). From the
350 maximum values obtained for the determined cations and anions, the following order of ionic
351 concentration can be established: $\text{Ca}^{2+} > \text{HCO}_3^-/\text{CO}_3^{2-} > \text{K}^+ > \text{Cl}^- > \text{Na}^+ > \text{Mg}^{2+} > \text{SO}_4^{2-}$.
352 Electric conductivity (EC) isovalues maps and the Σ cations within the district boundaries
353 were elaborated (Figs. 6A and 6B). The results of the ionic balance confirm a high cation
354 exchange capacity that occurs in clay and sandy loam soils with high concentration of Al^{3+}
355 [12], characteristics present in the Yellow Latosol Distrophic in association with
356 Hydromorphic Podzol and Concrete Lateritic present in the area of study. The most
357 important mineral sources found in the region are feldspars (plagioclase and potassic),
358 Muscovite and Biotite, and Mg^{2+} and Cl^- are all easily weathered, forming salts that are quite
359 soluble in the water and possibly being adsorbed (cations) by the clay particles. Considering
360 the PMV for the ionic composition, established by Ordinance N° 2914 of the Ministry of
361 Health [30], the waters of the wells monitored are in accordance with the pattern of
362 acceptance of consumption.

363 Considering the microbiological parameters determined by [16], all analyzed wells presented
364 some level of contamination by total coliforms. Concentrations of coliforms ranged from 21 to
365 585 NMP.100/mL, with sites 6 to 9 and 14 to 20 having values above 300 NMP.100/mL. This
366 pattern of contamination was also observed for fecal coliforms (*E. coli*), which ranged from 2
367 to 40 NMP.100/mL (mean 19 ± 12 NMP.100/mL). The presence of bacteria in the Coliform
368 group is indicative of contamination of the environment by fecal matter from warm-blooded
369 organisms. Contamination of groundwater by coliforms reveals a serious public health
370 problem, suggesting that sanitary quality is inadequate, indicating a risk situation for the
371 population that uses these waters. Based on current Brazilian legislation, which establishes
372 as a standard for human consumption the "total absence" in 100 mL of sample [30], the
373 indicative of coliforms, both total and fecal, in all wells sampled, even at low levels, is
374 alarming. According to [30], the presence of faecal coliforms in a 100 ml water sample leads
375 to non-compliance and invalidates the use of this resource. Isovalues maps of the
376 microbiologic parameters monitored in the wells were proposed (Figs. 6C and 6D).

377 Residential wells are generally poorly constructed, not obeying construction engineering
378 standards and techniques. Usually they are wells excavated manually, without any concern
379 for safety and hygiene, making them easily susceptible to microbiological contamination.
380 These wells are preferably found on the outskirts of the city, often close to pits, which cause
381 a serious problem for the preservation of water quality. According to [24], the most worrying
382 situation is when the pollutants go beyond the confining layers to reach the deep aquifers. In
383 these cases, the rupture of the rocky confinement makes possible the fecal contamination of
384 the aquifers considered less vulnerable.

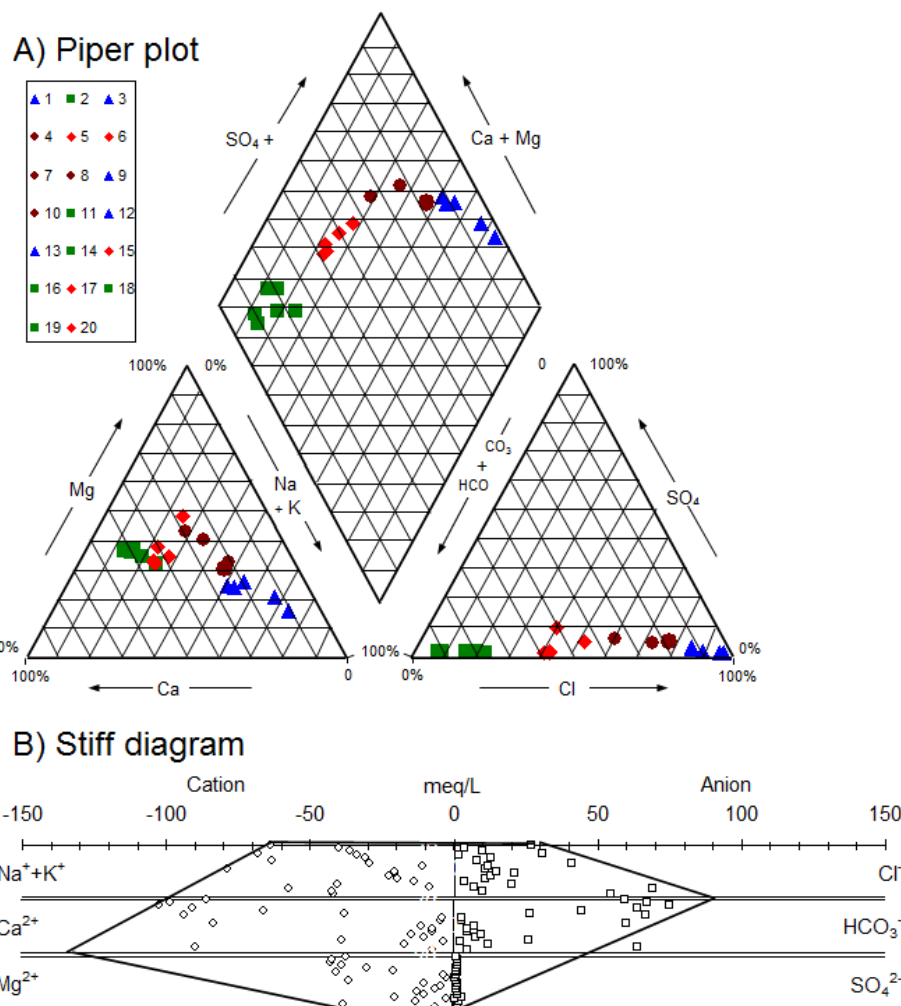


Fig. 5. A) Piper Diagram and B) Stiff diagram for the cations and anions determined in the waters of the 20 wells monitored, Municipality of Abaetetuba - PA.

4.3. Vulnerability Index

The study of aquifer vulnerability allows identifying areas more susceptible to contamination, due to the use and occupation of the soil and the capacity of the physical environment to provide some level of natural protection to the water system. This protection depends on the hydrogeological [34] and geological characteristics of the environment. The vulnerability to pollution of an aquifer will be different from one pollutant to another. The vulnerability assessment can be conducted for a specific contaminant, defined as 'Specific Vulnerability' or in general, for any contamination defined as 'Intrinsic Vulnerability' [35]. For this reason, vulnerability indexes include several hydrogeological parameters in their calculations. The results of the calculation of the DRASTIC Index [27], already multiplied by the respective weighting factors, were presented in isovalues map for the wells monitored (Fig. 7A). The DRASTIC Index ranged from 75 to 119, suggesting areas of 'moderate' to 'moderately-high' vulnerability. By definition, low vulnerability corresponds to the areas that present in the unsaturated zone a lithology composed of the mixture of clay, fine sand and silt, with water level above 25 meters. Moderate vulnerability suggests the presence of areas where

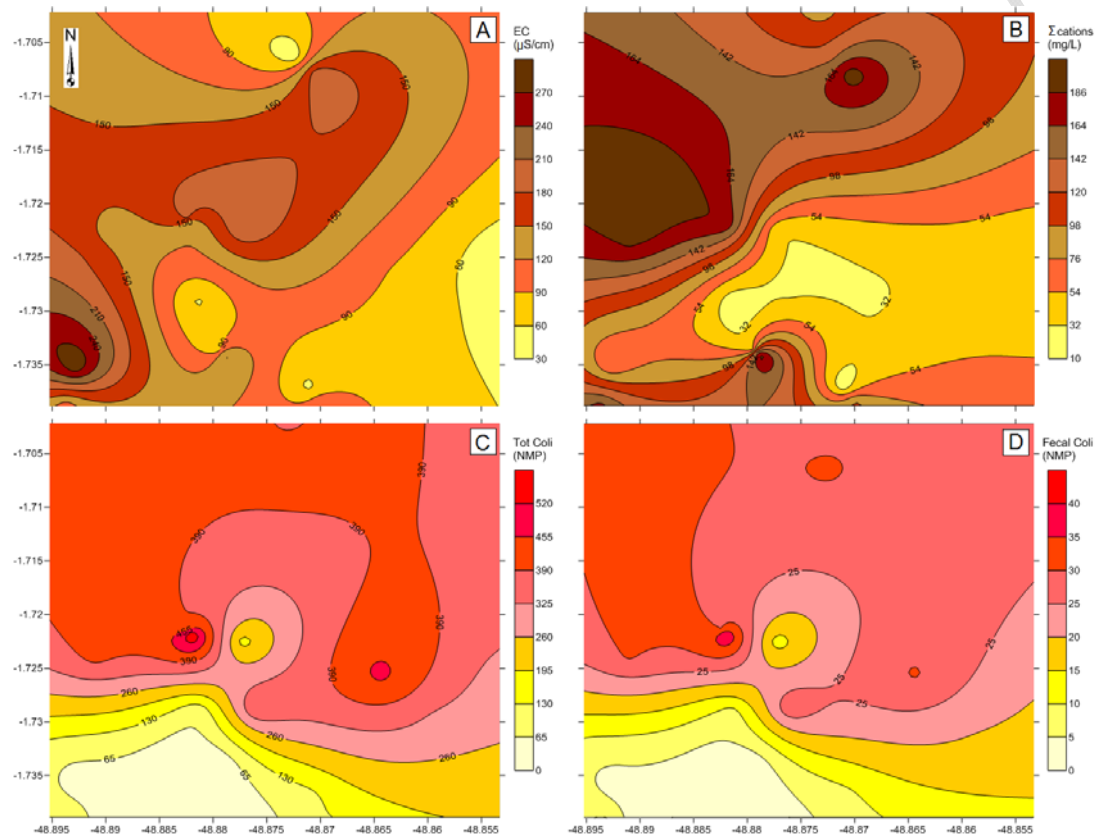
exploitable groundwater occurs (when the extraction overtakes the recharge), with a depth between 5 and 15 meters and underlies a material of medium to low permeability. In this case, the parameters depth to groundwater table (D), aquifer recharge (R) and soil media (S) were determinant for the classification of the risk of pollution of aquifers. The results of the calculation of the GOD Index [29] were presented in isovalues map for the wells monitored (Fig. 7B). The GOD Index ranged from 0.15 to 0.32, suggesting areas of 'low' to 'medium' vulnerability, thus confirming the trend observed by the DRASTIC Index. The comparison between DRASTIC vulnerability index and distribution patterns of the values of total solids dissolved (TDS) indicates areas where enhanced values of the TDS have been detected correspond with those with higher DRASTIC ratings. The categorized TDS map was showed as isovalues (Fig. 7C), and the TDS levels map and DRASTIC Vulnerability Index map were overlaid to obtain similarity. As result, using raster calculator and spatial analyst in ArcGis[®] 9.3 a raster map was generated in order to show the correlation between the two parameters (Fig. 7D), and the values corresponding to the data overlap oscillated within the range -3 to 3.

According to [36], the vulnerability to aquifer pollution can be defined as the sensitivity of groundwater quality to a pollutant load, based only on the intrinsic characteristics of the aquifer. Pollution and contamination of aquifers can occur in three ways: a) local sources - highly concentrated and reaching the aquifer at one point; b) linear sources - resulting from the infiltration of contaminated surface waters (streams and streams); and c) diffuse sources - in low concentration, however contaminating large areas because the pollutants are transported by rain, wind and agricultural activities. In the municipality of Abaetetuba, the main pollutant source observed was local. The lack of basic sanitation, especially sewage treatment, can generate several local sources, which continuously contaminate the wells, especially with fecal coliforms. Another local source that requires attention is the production of wastes from the washing and lubrication of vehicles (oils and greases), which has a high contaminant power. The use of septic tanks in large quantities in the region, precisely to compensate for the lack of sewage collection and treatment, can also cause pollution, in this case as a diffuse source.

Considering the local geological and hydrogeological aspects, and the fact that the Barreiras Group is predominant in the municipality, with partially deep and semi-confined aquifers, even if little thick, it can be said that the degree of vulnerability of the region is acceptable. In other words, the aquifer system in Abaetetuba is little vulnerable to contamination. The highest risk of contamination of the water is due to the wrong handling of the wells, in this case having superficial bacterial contamination. However, aquifers with zones of moderate vulnerability may, in the long term, undergo changes in the quality standard due to the presence of contaminants with mobility and effective persistence, such as metallic ions, hydrocarbons and poorly soluble salts. It should be understood that the results presented in this item are derived from the calculation of indices and, therefore, are subject to failure. Relevant information to identify areas considered potentially polluting, such as use and occupation of the land, are not included in the applied indexes. Thus, although the results were satisfactory, with low to moderately-high vulnerability ratings, a "state of alert" should be maintained for areas that indicated the presence of fecal coliforms and with a high population density. Preliminary studies based on the permeability and depth of the water table indicated the possibility of more vulnerable areas in the N-NE axis, towards the municipality of Bacarena.

The unsaturated layer, located in the upper part of the hydrogeological system, and the filtration capacity of the porous material that constitutes the aquifer, both exert important protection to groundwater quality, acting as a natural system of treatment of tailings, acting as a filter of the aquifers [37]. However, special attention should be given to recharge areas,

453 which can increase the degree of vulnerability of an aquifer through the presence of
 454 contaminants. The recharge sites may be susceptible to contamination, depending on the
 455 porosity and thickness of the sedimentary matrix that surrounds the aquifer. Over-
 456 exploitation of groundwater can also lead to serious environmental problems, such as
 457 reduction in the production capacity of wells; infiltration of low-quality groundwater from other
 458 more superficial aquifers; induction of lateral flows of brackish or saline water; and support
 459 loss of soil, resulting in stability problems of the built-up areas. Among the several
 460 applications, the results can help in the indication of areas susceptible to contamination;
 461 planning and land use; choice of suitable sites for new well drilling; and choice of locations
 462 for network installation to monitor and evaluate water contamination.



463

464 **Fig. 6. Mapas de isovalores para os parâmetros: electrical conductivity (EC, 6A); Σ**
 465 **cátions (6B); coliformes totais (6C) e fecais (6D) para os 20 poços monitorados no**
 466 **município de Abaetetuba (Pará, Brasil). Surfer® Golden Software, 9.11 (2010) and**
 467 **ArcGIS® 9.3 (2008) ESRI - USA.**

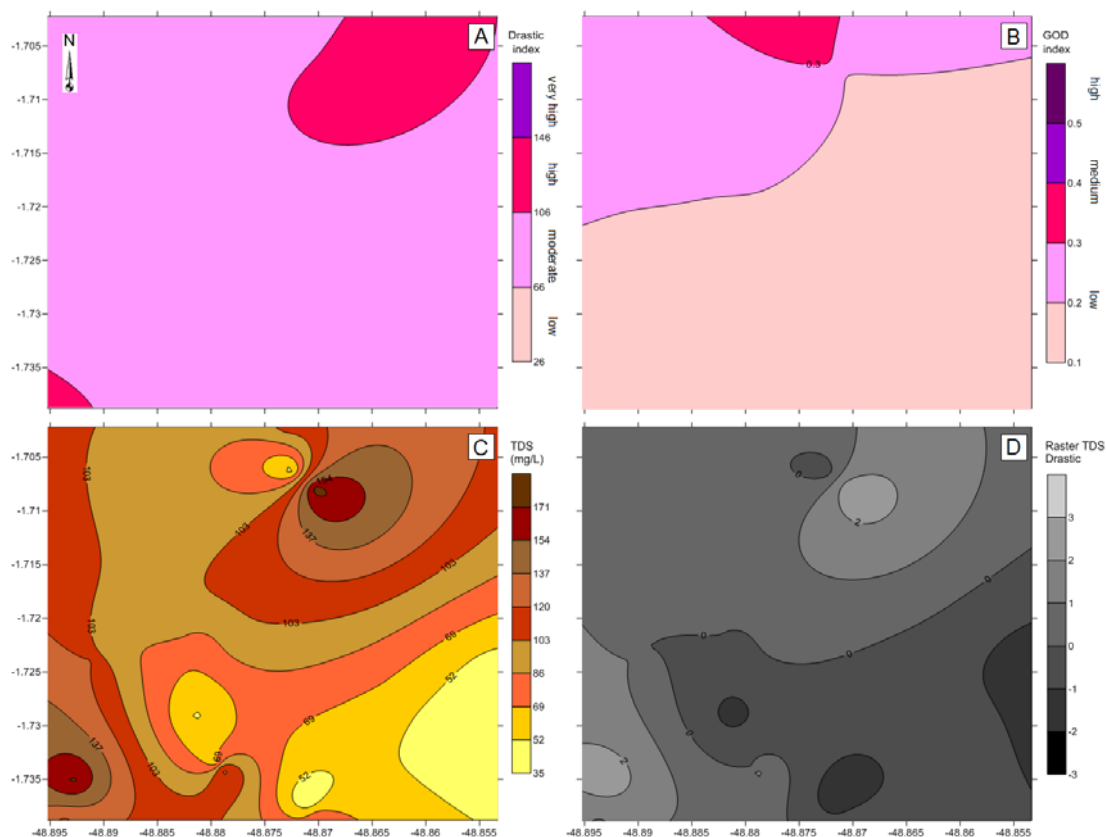


Fig. 7. Mapas de isovalores para os parâmetros: DRASTIC Vulnerability Index (7A); GOD Vulnerability Index (7B); TDS (mg/L) (7C) and correlation between categorized TDS and DRASTIC vulnerability maps (7D) para os 20 poços monitorados no município de Abaetetuba (Pará, Brasil). Surfer® Golden Software, 9.11 (2010) and ArcGis® 9.3 (2008) ESRI - USA.

5. CONCLUSION

The investigation of the geological aspects identified the stratigraphic units Barreiras, Post-Barrier Sediments and Recent Sediments (Quaternary) in the mesoregion where the municipality of Abaetetuba is located. The presence of the hydrogeological domains Barriers, Post-Barriers and Aluviões were also confirmed. The wells monitored are mostly belonging to the Barreiras aquifer, of medium to high depth and predominantly semi-confined with some porosity. The waters of the analyzed wells are mineralized, and their ionic balance, established by the Piper Diagram, suggested the presence of calcium chlorinated and calcium bicarbonate. The Stiff Diagram confirmed the results of the ionic balance. The results also confirm high cation exchange capacity, which occurs especially in clay-sandy and sandy- clay soils. The order of the ionic concentration for the 20 wells monitored was established as: $\text{Ca}^{2+} > \text{HCO}_3^-/\text{CO}_3^{2-} > \text{K}^+ > \text{Cl}^- > \text{Na}^+ > \text{Mg}^{2+} > \text{SO}_4^{2-}$. All wells showed some evidence of contamination by fecal coliforms (*E. coli*) remaining outside the standards of potability established by Ordinance N°. 2914 of the Ministry of Health [30]. The groundwater vulnerability maps, produced using the DRASTIC and GOD methods, suggested areas of 'low' to 'moderately-high' vulnerability, and the parameters depth to groundwater table, aquifer recharge and soil media were determinant for the classification of risk of pollution of aquifers. Despite this, the Barreiras aquifer in the semi-confined and free

492 areas may be becoming susceptible to microbiological contamination, mainly due to the
493 inadequate use and lack of maintenance of most wells. The intense exploitation of
494 groundwater, especially in areas of great population density, may be contributing to the
495 contamination of aquifers. For this reason, a 'state of alert' has been suggested for these
496 areas. The maps of isovalues and vulnerability indicated areas that require greater
497 environmental monitoring.

498 **COMPETING INTERESTS**

499 Authors have declared that no competing interests exist.

500 **AUTHORS' CONTRIBUTIONS**

501 All authors participated of the samples collection, data and statistical analysis and wrote the
502 first draft of the manuscript.

503 **CONSENT**

504 All the authors accepted the terms for publication, and we agree that, if the manuscript is
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510 **ETHICAL APPROVAL**

511 This section is not applicable in this manuscript.

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