

Influence of coral-reef complexity on species richness
and the hierarchical structuration of species abundances
in reef fish communities: A case study in south-east Brazil

Abstract

~~Increasing~~Growing complexity of coral habitat is expected to increase resource partitioning among co-occurring reef fish and, thereby, reduce to some extent the mean competitive intensity. ~~This will have, with associated or relative~~ consequences on the internal structuring of species in reef fish communities, in particular regarding species richness and ~~species evenness of~~ abundance ~~unevenness~~. Accumulating dedicated case studies are necessary, however, to get further empirical confirmations. The present ~~partial~~ analysis ~~aims to contribute in this respect, comparing~~ two coral-reef fish community settings that markedly differed in their degree of morphological complexity around two small islands ("Pai" island and "Mae" island) of Itaipu Sound, Brazil. ~~Available samplings remaining incomplete (as is often the case in practice),~~ ~~N~~umerical extrapolations were implemented, providing least-bias estimates for both total species richness and ~~the exhaustive~~ distribution of ~~species~~ abundances in ~~each~~ the two compared reef fish communities, based on a historical, although incomplete, sample set. ~~As expected, t~~he total species richness ~~is found to~~ increased in line with ~~higher greater degree level~~ of coral habitat complexity, while the species abundance unevenness ~~decreases~~. This decrease in abundance unevenness – reflecting the corresponding relaxation of the mean level of competitive intensity – is partly due to the direct, negative influence of species richness on abundance unevenness, as an overall trend. Beyond that, however, the relaxation is further strengthened by an additional “genuine” contribution – this time independent from the variation in species richness – and, as such, idiosyncratically attached to the improvement in habitat complexity.

Key-words: species diversity, ranked species abundance distribution, evenness, incomplete sampling, numerical extrapolation, Itaipu

1. INTRODUCTION

Coral reefs ~~as a whole~~, and the associated reef fish communities in particular, are ecosystems considered to exhibit embodying among the highest levels of diversity and biological complexity on Earth [1-6]. This high diversity is ~~still~~ enhanced by the close tight relationships that link binding reef fishes to the surrounding coral settings [27, 10]. ~~On the other hand~~ However, tropical marine ecosystems, especially those hosted by coral reefs, are under increasing threat, being particularly sensitive to ongoing anthropogenic impacts on the deteriorations of environment. This, in turn, necessitates constant urges to monitoring of the progressive change alteration of these ecosystems, especially focusing on the on-going reduction in species richness ~~as well~~ as and the increased unevenness of species abundances [4][3].

45 | Yet, detrimental anthropogenic activities are not the ~~unique-only~~ source of modification
46 | in the detailed structuring of species within animal communities. Other, non-
47 | anthropogenic ecological and environmental determinants may also be involved, whose
48 | own influences need to be assessed, in order to ~~be able to separate~~ *disentangle, in fine,*
49 | ~~what is~~ the genuine contribution of anthropogenic degradation ~~of to the~~ environment
50 | compartmentalization from ~~what is~~ the ~~mere~~ consequences of “natural” causes.
51 | Therefore, it is necessary, first, to improve our knowledge regarding the influence of
52 | different kinds of “natural” factors, such as, for example, the degree of structural
53 | complexity of coral setting [47-15].

54 |
55 | ~~Hence the necessity to collect as much data as possible on this subject by accumulating~~
56 | ~~dedicated case studies.~~

57 | To avoid making significant biased inferences regarding the main structural descriptors
58 | of ecological communities (such as total species richness and abundance unevenness), it
59 | is required to rely upon (sub-) exhaustive inventories [17-21]. Yet, incomplete
60 | samplings are almost unavoidable in practice, with species-rich communities having
61 | very uneven distribution of abundances, as is most often the case with reef fish
62 | communities. Hence, the need to complete the available partial samplings by
63 | implementing a reliable procedure of numerical extrapolation [22] that can provide
64 | least-biased estimates regarding the number of the unrecorded species, as well as the
65 | distribution of the abundances of these species. This is all the more important that rare
66 | species, that often escape recording in practice, may yet disproportionately contribute to
67 | the functional structuring of communities in the wild [23-33], the latter with numerous
68 | references therein. In particular, neglecting rare species can seriously reduce the
69 | capacity to detect ecological changes when analyzing species communities
70 | comparatively; thus “rare species are critical for bio-assessment” [33]. Fortunately, a
71 | recently developed procedure of numerical extrapolation takes into account these
72 | needs. Moreover, once properly numerically completed (and only when it is so [20]),
73 | the distribution of species abundances can provide synthetic data, in both qualitative
74 | and quantitative terms, about the underlying process that drives the hierarchical
75 | structuring of species abundances within community [34-38].

76 |
77 | The following issues (below), which form a significant and integral part of the results
78 | and discussion sections, should also be introduced in this section. The current shape of
79 | the Introduction section is: (1) not in line with the Results and Discussion sections, (2)
80 | insufficient in length for a full research paper (5-6 paragraphs).

81 |
82 | *Taxonomic dissimilarity between the two fish communities*

83 | *Testing for the type of process involved in the structuring of species abundances*

84 | *Beyond the rough abundance unevenness, the genuine intensity of the hierarchical*
85 | *structuring process*

86 | *Effect of habitat complexity on the true species richness and the taxonomic*
87 | *composition of associated fish communities*

88 | *Type of process involved in the structuring of species abundances*

89 *Effect of coral habitat complexity on the mean competitive intensity and the species*
90 *abundance unevenness within associated fish community*

91
92 **2. MATERIALS AND METHODS**

93 **2.1 - The reported field data**

94 The present study is based on two partial samplings of reef fish communities conducted
95 on two small islands ("Pai" island and "Mae" island) of Itaipu Sound, Niteroi, RJ, Brazil
96 (22°58'S - 43°02'W) ~~and reported in reference by Mendonça-Neto et al.[16]. All details~~
97 ~~regarding the precise locations of the compared habitats and the sampling procedure~~
98 ~~are provided in the open-access reference above by Mendonça-Neto et al. [16] and need~~
99 ~~not being repeated here.~~ An important point is that the numbers of individual
100 occurrences have been recorded for each species, thus making possible to implement
101 numerical extrapolations. These extrapolations are indeed required because the
102 subsistence of species recorded only once ("singletons") suggests that reported
103 samplings remain incomplete, as was indeed confirmed later. The number N_0 of collected
104 individuals (N_0) and the number R_0 of recorded species (R_0) in each of the two
105 communities are ~~given presented~~ in Table 1.

106 The coral habitat complexity, measured in terms of the chain link rugosity index
107 [11], ~~proves bein were~~ significantly higher at "Mae" island than at "Pai" island [16] (a third
108 reported community, at "Menina" island is not considered here, as it suffers intensive
109 fishing and important coastal runoff due to its proximity to the main shore [16]).
110

111 **2.2 - The Numerical Extrapolation procedure and its exploitation**

112 ~~2.3 To avoid making seriously biased inferences regarding the main structural~~
113 ~~descriptors of ecological communities (such as total species richness and~~
114 ~~abundance unevenness), it is required to rely upon (sub-) exhaustive inventories~~
115 ~~[17-21]. Yet, incomplete samplings are almost unavoidable in practice, with~~
116 ~~species-rich communities having very uneven distribution of abundances, as is~~
117 ~~most often the case with reef fish communities. Hence, the need to complete the~~
118 ~~available partial samplings by implementing a reliable procedure of numerical~~
119 ~~extrapolation [22] that can provide least-biased estimates regarding the number~~
120 ~~of the still unrecorded species, as well as the distribution of the abundances of~~
121 ~~these unrecorded species. This is all the more important that rare species, that~~
122 ~~often escape recording in practice, may yet disproportionately contribute to the~~
123 ~~functional structuring of communities in the wild: [23-33], the latter with~~
124 ~~numerous references therein. In particular, neglecting rare species can seriously~~
125 ~~reduce the capacity to detect ecological changes when analyzing species~~
126 ~~communities comparatively; thus "rare species are critical for bio-assessment"~~
127 ~~[33].~~

128 2.4.2.2 Fortunately, a recently developed procedure of numerical extrapolation
129 takes into account these needs (sections 2.2.1 to 2.2.3). Moreover, once properly
130 numerically completed (and *only* when it is so [20]), the distribution of species
131 abundances can provide synthetic data, in both *qualitative* and *quantitative*
132 terms, about the underlying process that drives the hierarchical structuring of
133 species abundances within community [34-38]. This is part of Introduction-
134 explaining the methodology VALUE = Introduction!(relocate)
135

136 2.2.1 - Implementation of the procedure of numerical extrapolation

137 * *Total species richness*: the least-biased estimation of the number of still undetected
138 species during partial sampling and the resulting estimation of the total species richness
139 of the partially sampled community are derived according to the procedure defined in
140 [39-40] and briefly summarized in Appendix 1, on the basis of the numbers f_x of species
141 observed x -times during partial sampling ($x = 1$ to 5). The same procedure allows to
142 derive the least-biased extrapolation of the “Species Accumulation Curve”, which
143 predicts the expected increase in the number of newly recorded species, $R(N)$, as a
144 function of the growing sampling size N (N : number of currently recorded individuals);
145 see Appendix 1 for computation. In practice, this extrapolation allows to *forecast* the
146 likely additional sampling efforts that would be required to obtain any desirable
147 increment in sampling completeness.

148 * *Species Abundance Distribution*: as mentioned above, the Species Abundance
149 Distribution (“S.A.D.”) is intended to provide the basic data necessary (i) to describe the
150 *pattern* of structuration of species abundances within community and (ii) to qualify and
151 quantify the underlying *process* that drives this structuration. Yet, to accurately exploit
152 its full potential [41, 42], the “S.A.D.” requires (i) to be *corrected* for the bias resulting
153 from drawing stochasticity during sampling of finite size and, still more importantly, (ii)
154 to be *completed* by *numerical extrapolation*, to the extent that sampling is suspected to
155 be incomplete, as revealed by the subsistence of singletons. The appropriate procedure
156 of correction and least-biased numerical extrapolation of the as-recorded partial “S.A.D.”
157 is described in details ~~in reference by Béguinot~~ [42], briefly summarized in Appendix 2
158 and concretely exemplified in ~~details-detail by Béguinot~~ [43]. Classically, the “S.A.D.” is
159 graphically presented with the (log-transformed) abundances a_i plotted against the rank
160 i of species, the latter being ordered by decreasing values of abundance (with, thus, a_1
161 and a_{S_t} respectively standing for the highest and the lowest abundances in a community
162 of S_t species).
163

164 2.2.2. *Abundance unevenness: the pattern of species abundance structuration*

165 Once numerically completed, the “S.A.D.” conveys all the relevant quantitative data
166 required to address the internal organization among species within a local community
167 [44]. In turn, the “S.A.D.” can be synthetically summarized by two of its major features:
168 the *total species richness* ‘ S_t ’ and the *degree ‘U’ of unevenness* of the abundance
169 distribution. Indeed, following Strong [45], it is the degree of *unevenness* – rather than
170 evenness – that should be preferred to address the hierarchical structuring of species
171 abundances in communities. According to the mode of representation of “S.A.D.”, it ~~goes~~
172 ~~is~~ natural to quantify the degree of abundance unevenness U as the average slope of the
173 log-transformed abundance decrease, as ~~already~~ proposed by Grzès [46], that is:

$$174 \quad U = [\log(a_1) - \log(a_{S_t})]/(S_t - 1) = [\log(a_1/a_{S_t})]/(S_t - 1) \quad (1)$$

175

176 with a_1 and a_{S_t} standing for the highest and the lowest abundances in the studied
177 community.

178 2.2.3. *Abundance unevenness: the underlying process of abundance structuration*

179 Beyond the unevenness pattern U , the underlying process of hierarchical structuration
180 of abundances is worth being considered, in terms of (i) the kind of *mechanism* involved,
181 and (ii) what determines the *intensity* of this structuring process, from which follows the
182 degree of abundance unevenness.
183

184 Very schematically, the kind of *mechanism* driving the hierarchical structuration of
 185 abundances may result either (i) from the major contribution of *one dominant* factor or
 186 (ii) from the combined contributions of *many mutually independent factors* acting
 187 together. This distinction can be tested by checking the conformity of the “S.A.D.” to
 188 either the *log-series* model or the *log-normal* model respectively [34, 47–50].

189 As regards now the *intensity* of the process of hierarchical structuration, it should be
 190 first emphasized that species richness has a *direct*, negative influence on abundance
 191 unevenness U, as a *general trend*, a point already highlighted by several authors [51–55].
 192 The likely underlying ecological origin of this overall trend (behind its “mathematical-
 193 like” appearance [51, 54]) is discussed in detail *in* by Béguinot [56]. Now, each particular
 194 community usually deviates more or less – often substantially – from this overall trend.
 195 So that it is appropriate to *consider and quantify separately*: (i) on the one hand, the
 196 contribution of this overall general trend and (ii) on the other hand, the more or less
 197 important deviation from this tendential influence, which specifically singularizes each
 198 particular community [55, 56]. As argued in detail *by* Béguinot [56], the direct, negative
 199 influence of species richness on abundance unevenness is adequately accounted for by
 200 the “broken-stick” theoretical distribution, originally conceptualized by MacArthur [57].
 201 Accordingly, it looks is relevant to standardize the “rough” abundance unevenness U to
 202 the corresponding rough abundance unevenness U’ of the “broken-stick” distribution,
 203 computed for the same species richness [58]. Doing so highlights to what extent the
 204 rough abundance unevenness U of a community actually deviates from the common
 205 overall trend, dictated by the tendential, direct influence of species richness [51, 52, 55,
 206 56, 58]. Accordingly, a *standardized unevenness index*, “I_{str}”, is defined by the ratio U/U’ [55,
 207 56]:
 208

$$I_{str} = U/U' = [\log(a_1/a_{st})/(S_t-1)]/[\log(a'_1/a'_{st})/(S_t-1)]$$

209 that is:

$$I_{str} = U/U' = \log(a_1/a_{st})/\log(a'_1/a'_{st}) \quad (2)$$

210 with a_1 and a_{st} standing for the highest and the lowest abundances in the studied
 211 community and a'_1 and a'_{st} standing for the highest and the lowest abundances in the
 212 corresponding “broken-stick” distribution computed for the same species richness S_t .

213 Thanks to this standardization – making it free from the direct influence of species
 214 richness – the index I_{str} allows for relevant, unbiased and meaningful comparisons
 215 between communities *differing by their species richness*, contrary to the rough
 216 abundance U, fully sensitive to this influence of species richness. In this respect, I_{str}
 217 deserves being considered as “genuinely” (idiosyncratically) attached to the
 218 corresponding community, independently of its particular species richness. Basically, the
 219 standardized abundance unevenness I_{str} satisfies the condition required in [53, 59]: “to
 220 *make sense*, (un)evenness must be independent of species richness”.

221 Now, from a *functional* point of view, the abundance unevenness U reflects the “mean
 222 competitive intensity” in the community (with “competitive intensity” being understood
 223 *sensu latissimo*, in its broadest scope, including both biotic and abiotic factors, as detailed
 224 *in* by Béguinot [56]). Accordingly, the standardized structuring index I_{str} reflects the mean
 225 competitive intensity, normalized (i.e. compared) to what it is in the broken-stick
 226 distribution at the same level of species richness. As the broken-stick model often fits
 227 rather well the structure of most bird communities [34, 57], it follows that the mean
 228 competitive intensity in a community is equal to I_{str} times that in *atypical bird community*
 229 *having the same species richness*. Thereby, the standardized structuring index I_{str} offers
 230
 231
 232

233 an *evocative benchmark* to appreciate more concretely the mean competitive intensity
 234 within community [51,–56]. And, of course, in its *functional* sense, as well as in its
 235 descriptive acceptance, the index I_{str} allows for relevant, unbiased and meaningful
 236 comparisons between communities, *regardless of their respective species richness*.

237 238 3. RESULTS

239 3.1 Estimated total species richness of each community

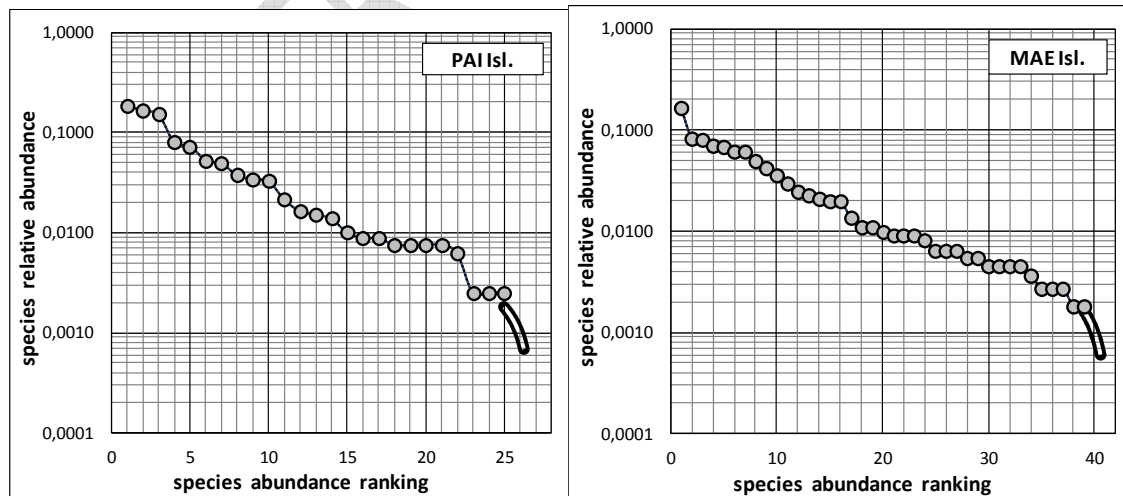
240 The two studied fish communities, at “Pai” and “Mae” islands, differ in their true (total)
 241 species richness, with *estimated* values $S_t = 26.2$ and $S_t = 40.5$, respectively (Table 1).
 242

243 **Table 1** – The number of collected individuals N_0 , the number of recorded species R_0 , the type of
 244 nonparametric estimator (Jackknife) selected as being the least-biased one, the estimated number Δ
 245 of unrecorded species, the resulting estimate of the “true” total species richness $S_t (= R_0 + \Delta)$, the
 246 resulting estimated level of sampling completeness R_0/S_t .

Reef fish community	PAI Isl.	MAE Isl.
nb. collected individuals N_0	770	1063
nb. recorded species $R_0 = R(N_0)$	25	39
selected least-biased estimator	JK-2	JK-2
number unrecorded species Δ	1.2	1.5
total species richness S_t	26.2	40.5
sample completeness R_0/S_t	95%	96%

247 248 3.2 Species Abundance Distributions numerically completed

249 The bias-corrected and numerically extrapolated *Species Abundance Distributions*
 250 (“S.A.D.”) of the two studied communities are provided in Figures 1 & 2. The abundances
 251 of the *recorded* species are plotted as *grey* circles, while the *extrapolated* part of the
 252 abundance distribution is plotted as a thick double line.
 253



254 **Figures 1 & 2** – The Species Abundance Distributions of reef fish communities at “Pai” island (left)
 255 and at “Mae” island (right). *Recorded*: discs; *numerically extrapolated* part: double line
 256
 257

258 3.3 Taxonomic dissimilarity between the two fish communities

259 3.3.1 Jaccard similarity index

260 Referring to recorded species lists, the fish community at “Pai” island is entirely nested
 261 taxonomically in the fish community at “Mae” island, the 25 recorded species in the
 262 former being all shared by the latter. Based on *recorded* data, the Jaccard Similarity
 263 Index is thus hypothetically evaluated as $J_r = 25/(25 + 39 - 25) = 0.64$. In turn,
 264 numerically extrapolated data ($S_i = 26.2$ and 40.5) allows to more surely specify that the
 265 actual Jaccard index is comprised between:

266 - at least, $J = 25/(26.2 + 40.5 - 25) = 0.60$, if no species are shared in common among
 267 the unrecorded species and

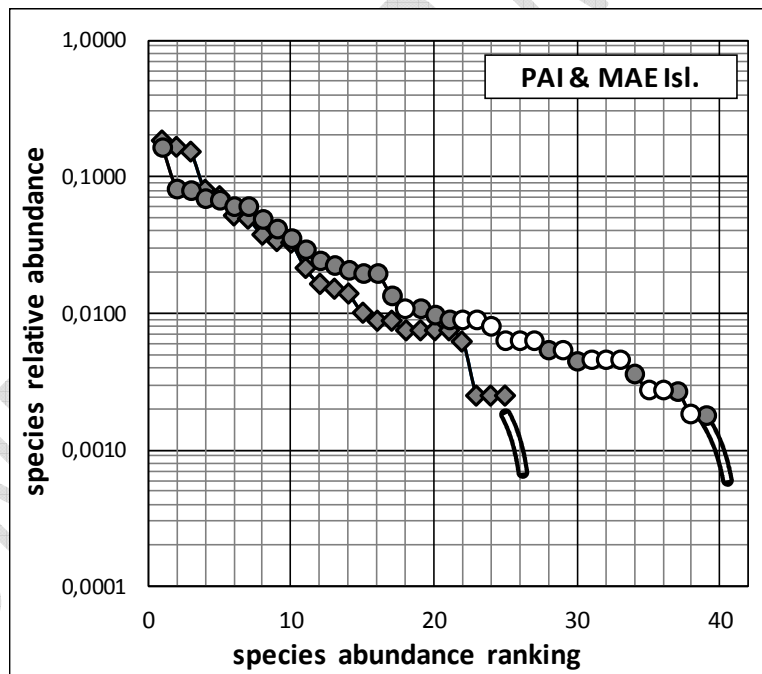
268 - at most, $J = 26.2/(26.2 + 40.5 - 26.2) = 0.65$, if unrecorded species in “Pai” island
 269 community is also shared by “Mae” island community.

270 That is –now more reliably based on numerically completed samplings – an estimated
 271 Jaccard similarity index comprised between 0.60 and 0.65.

272 3.3.2 Species exclusive to “Mae” community according to recorded data

273 Based on recorded data, 14 species (= 39–25) are considered exclusive to the
 274 community at “Mae” island and listed [in-byMendonça-Neto et al. \[16\]](#). [InterestinglyOf](#)
 275 [note is the fact that](#) –the average relative abundance of these 14 exclusive species is 5
 276 times lower than that of the 25 shared species (0.0071/0.0360). Although not
 277 surprising, this feature yet deserved being verified, as [graphically](#) highlighted in Figure
 278 3.

279



280

281 **Figure 3** –The Species Abundance Distributions of reef fish communities at “Mae” island (discs) and
 282 at “Pai” island (diamonds). Grey figures: species shared in common by both communities; white
 283 figures: species exclusive to the community at “Mae” island

284

285 [Report on Figures 4 \(*Chaetodonstriatus* Linnaeus 1758, common to both fish communities\), Figure 5](#)

286 [\(*Dactylopterusvolitans* \(Linnaeus 1758\), common to both fish communities\) and Figure 6](#)

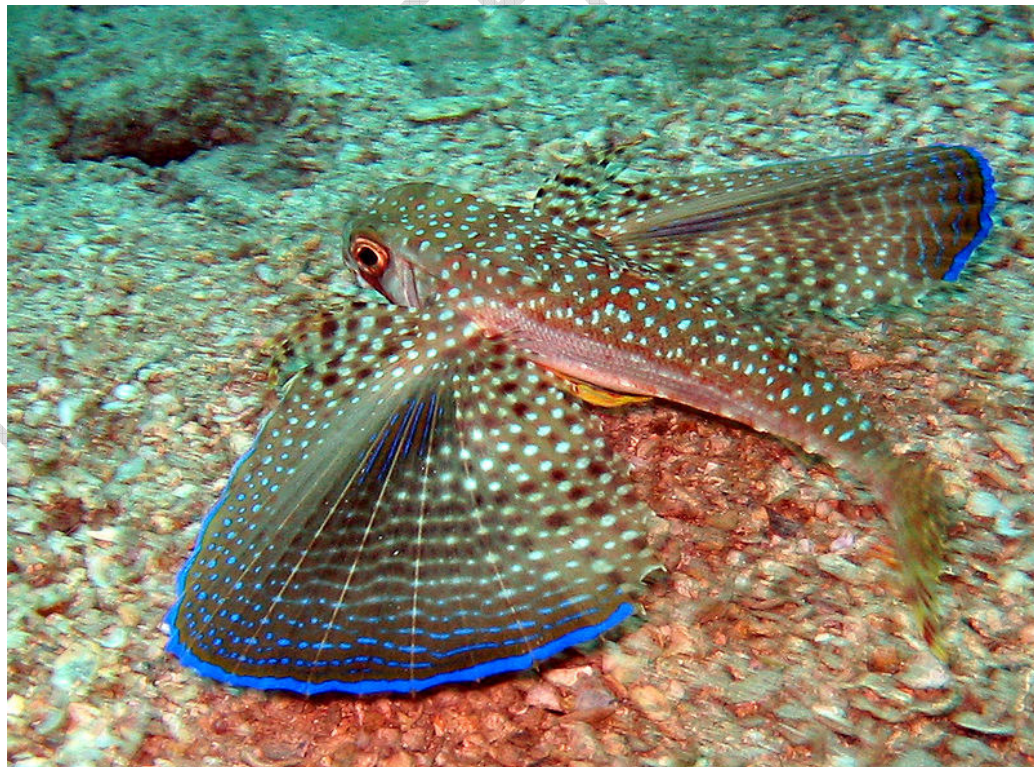
287 [\(*Pomacanthusparu* \(Bloch 1787\), recorded from Mae community only\)below to justify their inclusion.](#)

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Figure 4. *Chaetodonstriatus* Linnaeus 1758, common to both fish communities © Bernard E. Picton



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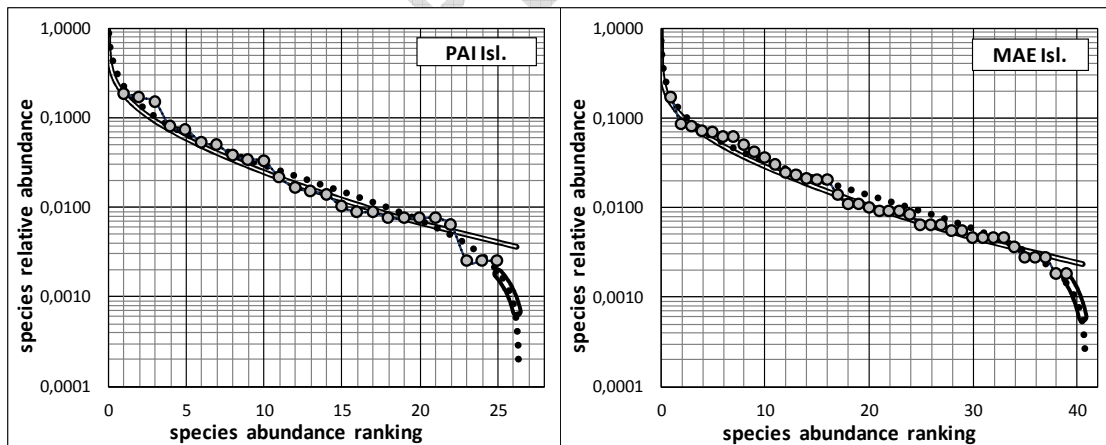
Figure 5. *Dactylopterusvolitans* (Linnaeus 1758), common to both fish communities © Carlos Henrique



298 **Figure 6.** *Pomacanthus paru* (Bloch 1787), recorded from Mae community only © Brian Gratwicke

3.4 Testing for the type of process involved in the structuring of species abundances

299 The numerically completed “S.A.D.s” of both studied communities clearly fit better the
 300 “log-normal” model than the “log-series” model (Figures 4 & 5). This remained rather
 301 uncertain as long as based exclusively upon recorded data and becomes quite clear only
 302 after considering the numerical extrapolation of abundance distributions.
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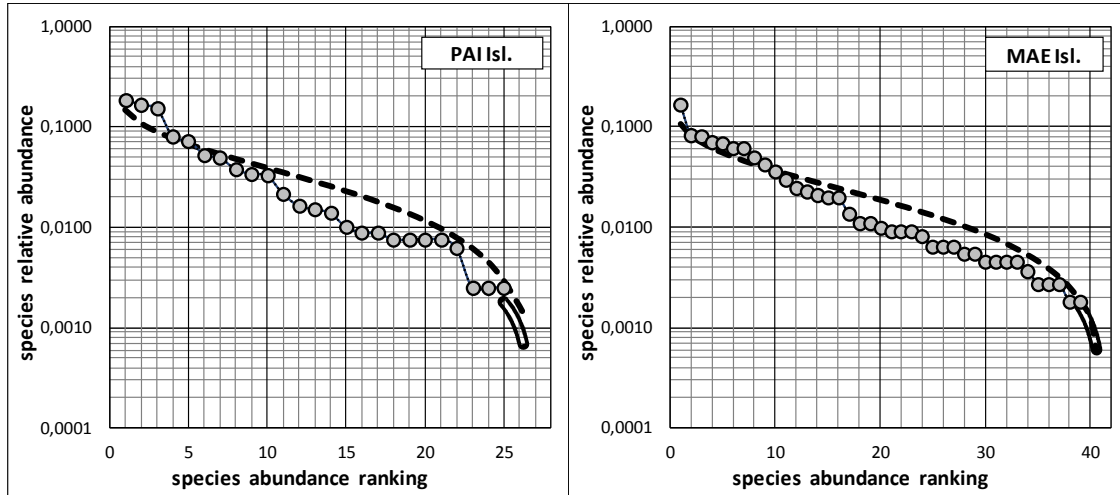


307 **Figures 4 & 5** – Two classical models: “log-normal” (coarse dotted line) and “log-series” (fine double
 308 line) compared to the numerically completed Species Abundance Distributions of each of the two
 309 studied communities. Best fit is with the “log-normal” distribution for both communities.
 310
 311

3.5 Beyond the rough abundance unevenness, the genuine intensity of the hierarchical structuring process

312 Figures 6, 7 and 8, allow to compare the average slope (U) of the “S.A.D.” to the average
 313 slope (U') of the corresponding “broken-stick” model (§ 2.2.2 & 2.2.3), from which is
 314
 315

316 derived the genuine intensity of the underlying structuring process $I_{str} = U/U'$ (equation
 317 (2)).
 318



319
 320 **Figures 6 & 7** – The Species Abundance Distribution for each studied community compared to the
 321 corresponding “broken-stick” model (dashed line) computed for the same species richness.
 322

323 The main results derived from this comparison are summarized synthetically in Table 2
 324 which highlights in particular the variations of the true total species richness S_t , the ratio
 325 a_1/a_{st} between the abundances of the commonest and rarest species, the rough
 326 abundance unevenness U and, finally, the standardized unevenness I_{str} .
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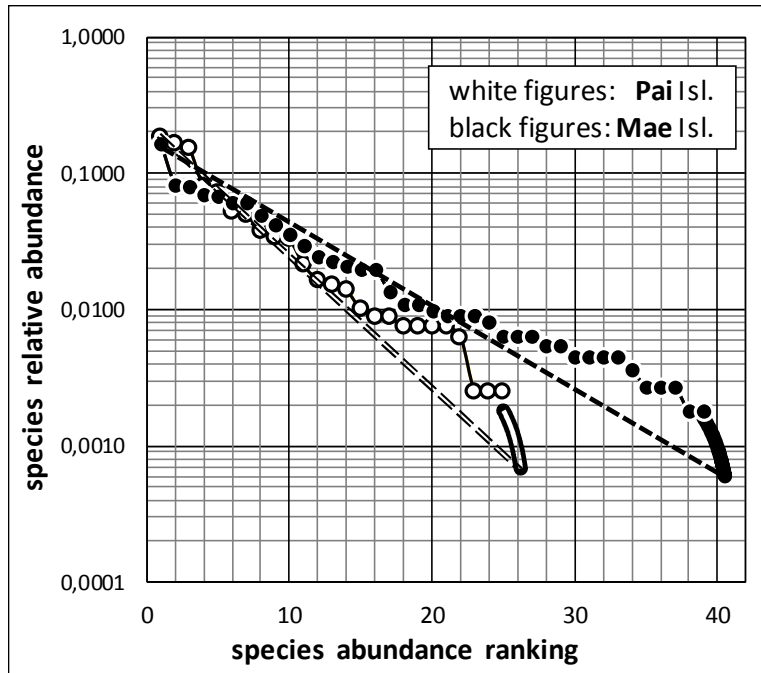
328 **Table 2** – A synthetic summary of the main quantitative features of the hierarchical organization of
 329 species abundances within community, as derived from numerically completed “S.A.D.s” : (i) the
 330 total species richness S_t of the community ; (ii) the relative abundances a_1 and a_{st} of the most and
 331 least abundant species (species rank 1 and S_t) ; (iii) the same, a'_1 and a'_{st} , for the “broken-stick”
 332 model, (iv) the rough unevenness of abundances in the community: $U = \log(a_1/a_{st})/(S_t-1)$; (v) the
 333 unevenness of abundances in the corresponding “broken-stick” distribution: $U' = \log(a'_1/a'_{st})/(S_t-1)$
 334 and, at last, (vi) the standardized unevenness index $I_{str} = U/U'$.

community	S_t	a_1	a_{st}	a_1/a_{st}	a'_1	a'_{st}	a'_1/a'_{st}	U	U'	I_{str}
PAI isld.	26.2	.1858	.000683	272	.1471	.00146	101	.0966	.0795	1.22
MAE isld.	40.5	.1657	.000599	277	.1063	.00061	174	.0615	.0565	1.09

335
 336 The numerically completed Species Abundance Distributions of both fish communities
 337 are plotted together in Figure 8, for a more straightforward appreciation of the effect of
 338 coral habitat complexity on the hierarchical structuration of species abundances. Rough
 339 abundance unevenness U is substantially less at “Mae” than at “Pai”, due to both:

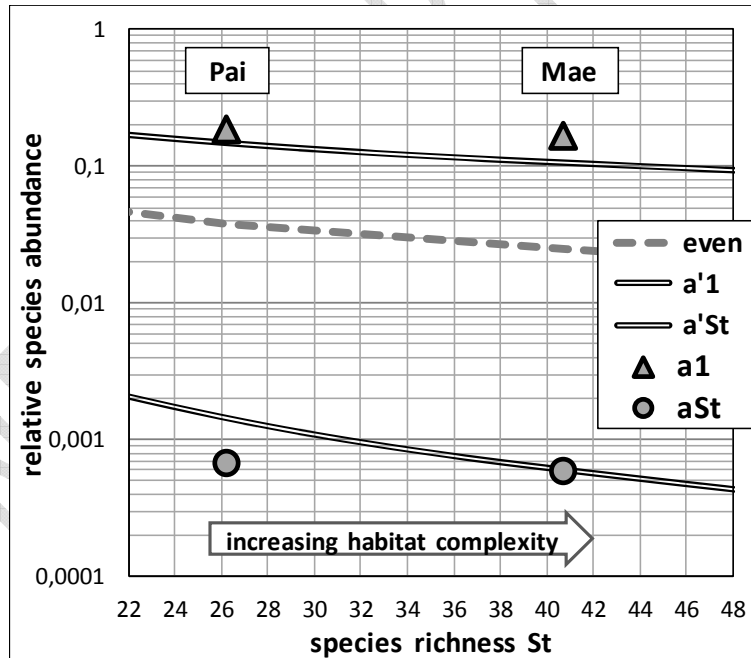
340 (i) the greater species richness at “Mae”, which negatively influence unevenness as a
 341 general overall trend and

342 (ii) the additional relaxation of the mean competitive intensity (leading to a
 343 corresponding additional reduction of abundance unevenness) allowed by the more
 344 complex habitat at “Mae” island, likely offering additional differentiated ecological
 345 niches and, thereby an easier resource partitioning among co-occurring species.
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Figure 8 – The Species Abundance Distributions of both fish communities plotted together to allow direct comparison of the influence of the coral habitat complexity, higher at “Mae” Island than at “Pai” island.



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Figure 9 – The maximum and minimum abundances, a_1 and a_{St} , for each of the two studied reef fish communities plotted jointly with the maximum and minimum abundances, a'_1 and a'_{St} of the “broken-stick” model and the uniform abundance level ($= 1/S_t$) of the perfectly “even” model. The range of species abundance (a_1/a_{St}) remains almost unchanged (+2%) from “Pai” to “Mae” while the corresponding range of species abundance in the broken-stick model (a'_1/a'_{St}) substantially increases (+73% for a_1/a_{St}) (cf. Table 2). This is at the origin of the decrease of I_{str} from “Pai” to “Mae” (Table 2).

360 | As a complement, Figure 9 highlights ~~graphically~~ how the highest and lowest
 361 abundances, a_1 and a_{st} , vary with increasing levels of species richness accompanying
 362 higher habitat complexity. Comparison is allowed with two theoretical referential
 363 models: the broken-stick distribution and the ideally even abundance distribution. Note
 364 that here, both a_1 and a_{st} , remain remarkably stable in spite of the substantial increase of
 365 species richness, contrasting in this with the two referential models. This stability
 366 emphasizes the efficacy of the relaxation of mean competitive intensity, likely allowed by
 367 the higher habitat complexity at “Mae”: this relaxation makes compatible a strong
 368 increase in species richness with a practically unchanged range of species abundances
 369 (i.e. without decreasing a_{st} nor increasing a_1).

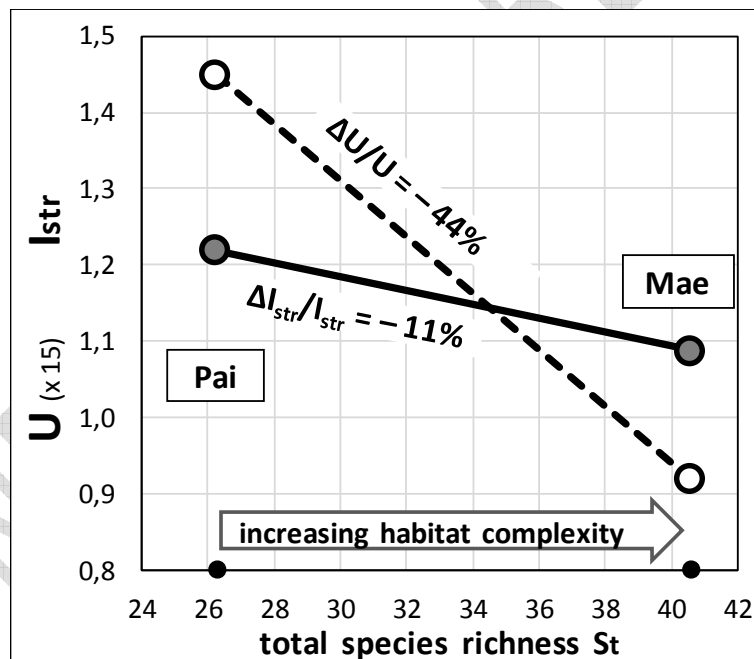
370 | ~~At last~~ Finally, the results in Table 2 highlight how each of the three structural
 371 parameters, S_t , U and I_{str} , respond respectively to the significant complexification of habitat
 372 at “Mae” as compared to “Pai”:

373 (i) the species richness S_t increases by $\Delta S_t/S_t = (S_{t2} - S_{t1}) / \frac{1}{2}(S_{t2} + S_{t1}) = 43\%$, which, in turn,
 374 corresponds to a variation $\Delta U/U = -33\%$ of the abundance unevenness for the broken-
 375 stick distribution (the latter accounting for the tendential negative influence of species
 376 richness on abundance unevenness: cf. section 2.2.3).

377 (ii) through abundance unevenness U decreases by $\Delta U/U = -44\%$

378 (iii) the standardized abundance unevenness I_{str} decreases by $\Delta I_{str}/I_{str} = -11\%$.

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380

381 **Figure 10** – The degree U of abundance unevenness (*dashed* line) and the intensity I_{str} of the
 382 underlying structuring process (*solid* line) plotted against the total species richness S_t , for the two
 383 studied communities at “Pai” and “Mae” Islands. The abundance unevenness U strongly decreases
 384 (by 44%) as the result of both (i) the tendential, negative direct influence of S_t upon U (contributing for
 385 33%) and (ii) a complementary, “genuine” contribution highlighted by the 11% decrease of the
 386 standardized unevenness I_{str} . Note that for commodity of graphical comparison between U and I_{str} , the
 387 values of U are uniformly multiplied by a same factor 15.
 388

389 As the degree of abundance unevenness accounts for the mean level of competitive
 390 intensity within community (cf. section 2.2.3), it follows that the complexification of the

391 coral habitat at “Mae” is conducive to a reduction of the mean competitive intensity in
392 the associated fish community attributable:

393 - for $\frac{3}{4}$ (= 33%/44%), to the direct tendential influence of the (43%) increase in
394 species richness;

395 - for $\frac{1}{4}$ (= 11%/44%), to an additional, *genuine* contribution to the relaxation of the
396 mean competitive intensity, specifically (“idiosyncratically”) attached to the community
397 at “Mae” (i.e. independently of the influence of its 43% higher species richness).

398 Figure 10 provides a graphical representation of these responses of S_t , U and I_{str} to the
399 complexification of habitat at “Mae”, as compared to “Pai”.

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4. DISCUSSION

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[[It has been argued [5, 16] that fish assemblages associated to tropical coral reefs exhibit close *positiverelationships* with the degree of habitat complexity [5, 16], resulting in particular from the morphology and the overall “rugosity” of the coral display [7-16, 60-63]. Coral species diversity has also been advocated has a determinant of richness and structuration of reef fish communities [60]. As a whole, a positive influence of habitat complexity on the species richness of reef fish communities has been highlighted.

Yet, in most case studies devoted to this subject, two important aspects, conditioning the relevance of the analysis, had been neglected and, therefore still needed to be addressed:

- the bias resulting from the (often unavoidable) *incompleteness* of available samplings [9, 10, 43, 64-67];

- beyond its role on species richness, the usually overlooked effect of habitat complexity on the distribution of species abundance, especially the degree of *abundance unevenness*.

Here, as in preceding reports dealing with coral reef-associated communities [9, 10, 43, 64-67], *numerical* extrapolations are implemented to compensate for the lack of exhaustive samplings, thus providing least-biased estimates of the number of unrecorded species and their respective abundances – using a dedicated procedure, recently made available [39, 42]. Thereby, the full-range of the Species Abundance Distribution is derived, including the set of species that had remained undetected. In particular, major traits of community organization – the true (total) species richness S_t , the degree of species abundance unevenness U and the standardized abundance unevenness I_{str} – are provided in Tables 1 & 2 and Figures 1 & 2.

- This look like introduction as it does not directly discuss the obtained results – relocate to Introduction section (or remove?) – Immediately start discussion obtained results – all else goes to relevant sections (Introduction or methodology).]

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4.1 - Effect of habitat complexity on the true species richness and the taxonomic composition of associated fish communities

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A higher species richness in reef fish community is anticipated at “Mae” as compared to “Pai”, answering a significantly more tormented coral habitat (10% higher rugosity); what was confirmed with a 43% higher estimated true species richness at “Mae”.

The substantial taxonomic dissimilarity highlighted between the two fish communities (estimate of Jaccard similarity index comprised between 0.60 and 0.65) results from the community at “Pai” being essentially nested in the community at “Mae” (rather than from simple taxonomical turn-over). The community at “Pai” only keeps the subset of

437 the most abundant species from “Mae”, being deprived from the rarer species which,
438 apparently, can only survive ~~thanks~~ due to the more tormented coral habitat at
439 “Mae” (Figure 3). This fits the likely expected trend according to which depauperate
440 ecological communities tend to lose their less abundant species first and, similarly, the
441 trend according to which taxonomic turnover, if any, tends to preferentially target the
442 less abundant species first [67].

443 **4.2 - Type of process involved in the structuring of species abundances**

444 The numerically-completed “S.A.D.s” of both fish communities clearly fit best the “log-
445 normal” than the “log-series” models (Figures 4 and 5), thereby suggesting that the
446 hierarchical structuring of species abundances is governed by the combined influence of
447 many independent factors, rather than by one or very few dominant factor(s). Note that
448 the conclusion would have remained less clear if only the recorded part of the “S.A.D.s”
449 was available, thereby emphasizing, once again, the interest of numerical extrapolations
450 of incomplete inventories. The conformity of both “S.A.D.s” to the “log-normal” model, is
451 not surprising, being rather common in most species-rich communities, at least when
452 they are not subjected to excessively harsh environmental stresses (pollutions, etc...) [47,
453 49, 50].

455 **4.3 - Effect of coral habitat complexity on the mean competitive intensity and the species abundance unevenness within associated fish community**

456 The lower habitat complexity at “Pai” is associated to a stronger level of rough
457 abundance unevenness U (Table 2, Figure 10), reflecting in turn a substantial increase in
458 mean competitive intensity resulting from the more uniform habitat at “Pai”. The latter
459 likely offers less varied feeding niches and fewer opportunities of protection against
460 potential predation. Hence the increase in mean competitive intensity at “Pai” and the
461 parallel decrease in species richness, already underlined above. Moreover, the increase
462 of competitive intensity and abundance unevenness U slightly exceeds what is expected
463 from the negative direct relationship between U and S_t (section 2.2.3): the standardized
464 unevenness I_{str} (made freed from the direct influence of species richness) remains 11%
465 larger at “Pai” than it is at “Mae” (Table 2, Figure 10). This clearly underlines the role of
466 “physical” simplification in the coral habitat upon the degree of mean competitive
467 intensity and the resulting severity in the hierarchical structuring of species abundances
468 in associated fish community. Conversely, a higher physical complexity of the coral
469 habitat likely allows an improved resource partitioning (*sensu latissimo*) among co-
470 occurring fish species and, accordingly, leads to both an improved relaxation of mean
471 competitive intensity (hence the decrease of both U and I_{str}) and the resulting
472 opportunity to accommodate a larger number S_t of co-occurring species.

475 **4.4 - Comparison with other case studies dealing with the same subject**

476 The influence of coral habitat complexity on the *total species richness* S_t , the *rough*
477 *abundance unevenness* U , the *standardized abundance unevenness* I_{str} and the *degree of*
478 *taxonomic differentiation*, are compared between three reef fish communities
479 respectively located at Bonaire (Dutch Caribbean) [9], Gorgona Island (Columbia) [10]
480 and Itaipu (present study): Table 3.

481 **1) total species richness S_t**

482 The positive role of higher coral habitat complexity on the total species richness of the
483 associated fish community was also reported at Gorgona Island, while (rather

484 surprisingly) the studied site at Bonaire provides what seems a counterexample, with a
 485 slight decrease in species richness associated to an apparently stronger habitat
 486 complexity – which remains unexplained.

487 *2) rough abundance unevenness U*

488 Due to the prominent negative direct influence of species richness on abundance
 489 unevenness, the answer of rough abundance unevenness U to habitat complexity is
 490 systematically opposite to the answer of species richness. Accordingly, the rough
 491 unevenness decreases at Itaipu and Gorgona, but slightly increases at Bonaire.

492 *3) standardized abundance unevenness I_{str}*

493 The systematic decrease, in all three case studies, of the standardized unevenness
 494 I_{str} with higher habitat complexity is still more interesting, and ecologically significant, as
 495 this reflects the *likely systematic* contribution of higher habitat complexity to the
 496 reduction of the mean competitive intensity (beyond the general trend linking positively
 497 relaxed competition and species richness).

498 *4) Taxonomic differentiation: Jaccard similarity index J*

499 Jaccard similarity between the two compared communities at Itaipu is estimated in the
 500 range 0.62 ± 0.03 , ~~that~~ *which* is close to the estimated values of Jaccard index at Gorgona
 501 and Bonaire. Some consistency is thus highlighted also regarding the degree of
 502 taxonomic differentiation induced by increased habitat complexity, with Jaccard index
 503 remaining around $J = 0.60$ to 0.65 .

504
 505 **Table 3** –The consequences of an increase in habitat complexity – compared in three sites (Brazil,
 506 Columbia, Dutch Caribbean) – in terms of: (i) variation $\Delta S_t/S_t$ of total species richness, (ii) variation
 507 $\Delta U/U$ of rough abundance unevenness and its contributors, $\Delta U'/U'$ and $\Delta I_{str}/I_{str}$ (with $\Delta U/U = \Delta U'/U' +$
 508 $\Delta I_{str}/I_{str}$). Also mentioned is the estimated range for the Jaccard similarity index between compared
 509 communities.

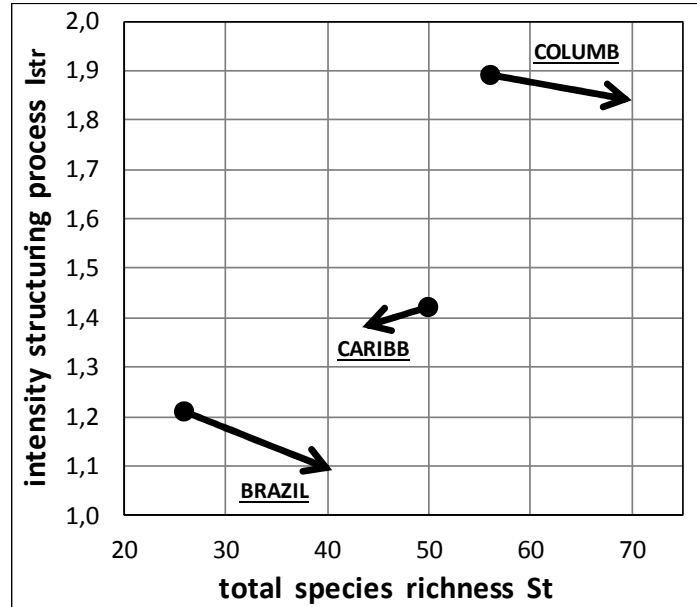
investigated sites (each of them including two communities differing by their degree of habitat complexity)	species richness S_t	rough mean competitive intensity U	tendential influence from S_t	additional genuine contribution	level of taxonomic proximity
	$\Delta S_t/S_t$	$\Delta U/U$	$\Delta U'/U'$	$\Delta I_{str}/I_{str}$	Jaccard index
Brazil – Itaipu Isld.	+ 43 %	– 44 %	– 33 %	– 11 %	$0.62 \pm$ 0.03
Columbia Gorgona Isld.	+ 22 %	– 21 %	– 18 %	– 3 %	$0.65 \pm$ 0.15
Caribbean Bonaire Isld.	– 14 %	+7 %	+10 %	– 3 %	$0.61 \pm$ 0.24
<i>average trend</i>	+ 17 %	– 19 %	– 13 %	– 6 %	0.63

510

511 *5) synthetic view*

512 The four main descriptors of the internal structuring in a community (S_t , U, U', I_{str}) are
 513 linked by two relationships: equation (2) above, which relies U, U', I_{str} and the
 514 mathematical dependence of broken-stick unevenness U' upon S_t (equation (2) in [56]).
 515 Accordingly, two descriptors only, among the four, can be chosen as mutually
 516 independent – and, therefore, two descriptors are sufficient to summarize the
 517 information conveyed by the all four descriptors. Here, *total species richness* S_t and
 518 *standardized unevenness* I_{str} are selected as the two independent descriptors. Figure

519 11 provides a synthetic view of how increased coral-habitat complexity affects these two
 520 descriptors, within each of the three reef fish communities considered above.
 521 In particular, the systematic decrease of I_{str} highlights the “genuine”, idiosyncratic
 522 contribution of higher habitat complexity to the relaxation of the mean competitive
 523 intensity, in complement to its indirect contribution to relaxation, *via* species richness.
 524



525
 526 **Figure 11** – The consequence of higher habitat complexity on (i) the total species richness S_t and
 527 (ii) the standardized unevenness I_{str} , at three sites (Brazil, Columbia, Dutch Caribbean).
 528

529 5. CONCLUSION

530 Relevant data remains scarce regarding the influence of coral habitat complexity on the
 531 internal structuring of associated reef fish communities, as this requires first:

- 532 - dealing with either exhaustive or duly numerically completed samplings of fish
 533 communities,
- 534 - considering not only the influence of habitat complexity on total species richness but
 535 also species abundance unevenness in associated fish communities.

536 To my knowledge, only three such case studies, summarized in Table 3 (including the
 537 present report), are presently available, which, obviously, makes it rather difficult to
 538 draw well-founded conclusions. Accordingly, the following proposals are only forward-
 539 looking suggestions that require additional empirical confirmations, although they seem
 540 fairly consistent with reasonable expectations.

541 Increased physical complexity of coral habitat (often associated to, or resulting from
 542 higher taxonomic diversity within coral settings) is expected to offer more diversified
 543 feeding niches as well as more diversified protective shelters against predators. In
 544 short, more opportunities to improve “resource partitioning” among co-occurring fish
 545 species. This improved partitioning of the available resource is expected to allow for a
 546 larger number of species to share the same habitat, thus resulting in a substantial
 547 relaxation of mean competitive intensity and, consequently, a reduction in the level of
 548 abundance unevenness. Indeed, this parallel tendency for both a relaxed competitive
 549 intensity (reflected in the substantial decrease of rough abundance unevenness) and a
 550 growing total species richness is supported by the presently available data, at least as an

551 average trend. Moreover, an additional, genuine contribution to the relaxation of the
 552 mean competitive intensity (reflected by the decrease of the standardized unevenness
 553 index I_{str}) is systematically highlighted.
 554 Once again subjected to further confirmation, these results provide new empirical
 555 support regarding the benefits that reef fish communities can derive from more complex
 556 coral habitat: higher species richness and improved stability, favored by further
 557 relaxation of the mean competitive intensity.

558
 559

560 Appendix 1

561 Bias-reduced extrapolation of the Species Accumulation Curve and associated 562 estimation of the number of missing species, based on the recorded numbers of 563 species occurring 1 to 5 times

564 Consider the survey of an assemblage of species of size N_0 (with sampling effort
 565 N_0 typically identified either to the number of recorded individuals or to the number of
 566 sampled sites, according to the inventory being in terms of either species abundances or
 567 species incidences), including $R(N_0)$ species among which f_1, f_2, f_3, f_4, f_5 , of them are
 568 recorded 1, 2, 3, 4, 5 times respectively. The following procedure, designed to select the
 569 less-biased solution, results from a general mathematical relationship that constrains
 570 the theoretical expression of *any* theoretical Species Accumulation Curves $R(N)$ [see [39,
 571 68, 69]:

$$572 \frac{\partial^x R(N)}{\partial N^x} = (-1)^{(x-1)} f_{x(N)} / C_{N,x} \approx (-1)^{(x-1)} (x! / N^x) f_{x(N)} \quad (\approx \text{as } N \gg x) \quad (A1.1)$$

573 Compliance with the mathematical constraint (equation (A.1)) warrants *reduced-bias*
 574 expression for the extrapolation of the Species Accumulation Curves $R(N)$ (i.e. for N
 575 $> N_0$). Below are provided, accordingly, the polynomial solutions $R_x(N)$ that respectively
 576 satisfy the mathematical constraint (A1.1), considering increasing orders x of derivation
 577 $\partial^x R(N) / \partial N^x$. Each solution $R_x(N)$ is appropriate for a given range of values of f_1
 578 compared to the other numbers f_x , according to [39]:

580
 581
 582

$$583 \text{ * for } f_1 \text{ up to } f_2 \rightarrow R_1(N) = (R(N_0) + f_1) - f_1 \cdot N_0 / N$$

584

$$585 \text{ * for larger } f_1 \text{ up to } 2f_2 - f_3 \rightarrow R_2(N) = (R(N_0) + 2f_1 - f_2) - (3f_1 - 2f_2) \cdot N_0 / N -$$

$$586 (f_2 - f_1) \cdot N_0^2 / N^2$$

587

$$588 \text{ * for larger } f_1 \text{ up to } 3f_2 - 3f_3 + f_4 \rightarrow R_3(N) = (R(N_0) + 3f_1 - 3f_2 + f_3) - (6f_1 - 8f_2 + 3f_3) \cdot N_0 / N -$$

$$589 (-4f_1 + 7f_2 - 3f_3) \cdot N_0^2 / N^2 - (f_1 - 2f_2 + f_3) \cdot N_0^3 / N^3$$

590

$$591 \text{ * for larger } f_1 \text{ up to } 4f_2 - 6f_3 + 4f_4 - f_5 \rightarrow R_4(N) = (R(N_0) + 4f_1 - 6f_2 + 4f_3 - f_4) -$$

$$592 (10f_1 - 20f_2 + 15f_3 - 4f_4) \cdot N_0 / N - (-10f_1 + 25f_2 - 21f_3 + 6f_4) \cdot N_0^2 / N^2 -$$

$$593 (5f_1 - 14f_2 + 13f_3 - 4f_4) \cdot N_0^3 / N^3 - (-f_1 + 3f_2 - 3f_3 + f_4) \cdot N_0^4 / N^4$$

594

$$595 \text{ * for } f_1 \text{ larger than } 4f_2 - 6f_3 + 4f_4 - f_5 \rightarrow R_5(N) = (R(N_0) + 5f_1 - 10f_2 + 10f_3 - 5f_4 + f_5)$$

$$596 - (15f_1 - 40f_2 + 45f_3 - 24f_4 + 5f_5) \cdot N_0 / N - (-20f_1 + 65f_2 - 81f_3 + 46f_4 - 10f_5) \cdot N_0^2 / N^2 -$$

$$597 (15f_1 - 54f_2 + 73f_3 - 44f_4 + 10f_5) \cdot N_0^3 / N^3 - (-6f_1 + 23f_2 - 33f_3 + 21f_4 - 5f_5) \cdot N_0^4 / N^4 -$$

$$598 (f_1 - 4f_2 + 6f_3 - 4f_4 + f_5) \cdot N_0^5 / N^5$$

599

600 The associated non-parametric estimators of the number Δ_j of missing species in the
 601 sample [with $\Delta_j = R(N=\infty) - R(N_0)$] are derived immediately:
 602
 603

604 * $f_1 \leq f_2 \rightarrow \Delta_{j1} = f_1 ; R_1(N)$
 605

606 * $f_2 < f_1 \leq 2f_2 - f_3 \rightarrow \Delta_{j2} = 2f_1 - f_2 ; R_2(N)$
 607

608 * $2f_2 - f_3 < f_1 \leq 3f_2 - 3f_3 + f_4 \rightarrow \Delta_{j3} = 3f_1 - 3f_2 + f_3 ; R_3(N)$
 609

610 * $3f_2 - 3f_3 + f_4 < f_1 \leq 4f_2 - 6f_3 + 4f_4 - f_5 \rightarrow \Delta_{j4} = 4f_1 - 6f_2 + 4f_3 - f_4 ; R_4(N)$
 611

612 * $f_1 > 4f_2 - 6f_3 + 4f_4 - f_5 \rightarrow \Delta_{j5} = 5f_1 - 10f_2 + 10f_3 - 5f_4 + f_5 ; R_5(N)$
 613

614 **N.B.1:** As indicated above (and demonstrated in details in [39]), this series of
 615 inequalities define the ranges that are best appropriate, respectively, to the use of each
 616 of the five estimators, JK-1 to JK-5. That is the respective ranges within which each
 617 estimator will benefit of minimal bias for the predicted number of missing species.

618 Besides, it is easy to verify that another consequence of these preferred ranges is that
 619 the selected estimator will *always* provide the highest estimate, as compared to the
 620 other estimators. Interestingly, this mathematical consequence, of general relevance, is
 621 in line with the already admitted opinion that all non-parametric estimators provide
 622 *under-estimates* of the true number of missing species [19, 21, 70-72]. Also, this shows
 623 that the approach initially proposed by [73] – which has regrettably suffered from its
 624 somewhat difficult implementation in practice – might be advantageously reconsidered,
 625 now, in light of the very simple selection key above, of *far much easier practical use*.
 626

627 **N.B.2:** In order to reduce the influence of drawing stochasticity on the values of the f_x ,
 628 the as-recorded distribution of the f_x should preferably be smoothened: this may be
 629 obtained either by rarefaction processing or by regression of the as-recorded
 630 distribution of the f_x versus x .
 631

632 **N.B. 3:** For f_1 falling beneath $0.6 \times f_2$ (that is when sampling completeness closely
 633 approaches exhaustivity), then Chao estimator may alternatively be selected: see
 634 reference [40].
 635

636 Appendix 2

637 ***Correction and extrapolation (when required) of the as-recorded S.A.D.***

638 N.B.: details regarding the derivation of the following expressions are provided in [42].

639 1) *Correction for bias of the recorded part of the S.A.D.*

640 The bias-corrected expression of the true abundance, \tilde{a}_i , of species of rank 'i' in the S.A.D.
 641 is given by:

642
$$\tilde{a}_i = p_i \cdot (1 + 1/n_i) / (1 + R_0/N_0) \cdot (1 - f_1/N_0) \quad (A2.1)$$

643 where N_0 is the actually achieved sample size, $R_0 (=R(N_0))$ the number of recorded
 644 species, among which a number f_1 are singletons (species recorded only once), n_i is the
 645 number of recorded individuals of species 'i', so that $p_i = n_i/N_0$ is the recorded frequency
 646 of occurrence of species 'i', in the sample. The crude recorded part of the "S.A.D." –
 647 expressed in terms of the series of as-recorded frequencies $p_i = n_i/N_0$ – should then be
 648 replaced by the corresponding series of expected true abundances, \tilde{a}_i , according to
 649 equation (A2.1).

650 2) *Extrapolation of the recorded part of the S.A.D. accounting for the complementary*
651 *abundance distribution of the set of unrecorded species*
652 The following expression stands for the estimated abundance, a_i , of the unrecorded
653 species of rank i (thus for $i > R_0$):

$$654 \quad a_i = (2/N_i)/(1+ R(N_i)/N_i).(1- [\partial R(N)/\partial N]_{N_i}) \quad (A2.2)$$

655 which, in practice, comes down to: $a_i \approx (2/N_i)/(1+ R(N_i)/N_i)$, as $f_1(N)$ already becomes
656 quite negligible as compared to N for the extrapolated part.

657 This equation provides the extrapolated distribution of the species abundances a_i (for
658 $i > R(N_0)$) as a function of the least-biased expression for the extrapolation of the species
659 accumulation curve $R(N)$ (for $N > N_0$), 'i' being equal to $R(N_i)$. The key to select the least-
660 biased expression of $R(N)$ is provided at Appendix 1.

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662

663 **COMPETING INTERESTS**

664 Author has declared that no competing interests exist.

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