

Multiple Degradation and Resistance Capabilities of Marine Bacteria Isolated from Niger Delta, Nigeria on Petroleum Pollutants and Heavy Metals

ABSTRACT

Aims: To determine multiple degradation and resistance capabilities of marine bacteria isolated from Rivers State, Nigeria on petroleum pollutants and heavy metals.

Study Design: Nine treatments and the controls designs were set up in triplicates containing 100 mL of sterile modified mineral basal medium in 250 mL conical flasks supplemented with 50, 100, 200 and 300 ppm of xylene, anthracene and pyrene each; 1 % of other petroleum pollutants and 300 ppm of heavy metals, nine marine hydrocarbon degraders and incubated at 24 °C for 5 - 7 days. The nine treatments and control set ups designated as ANT1, XYL2, PYR3, ANT4, PYR5, ANT6, XYL7, XYL8, PYR9 and CTRL (Without hydrocarbons) were used to determine the multiple degradability of the marine bacteria.

Place and Duration of Study: Department of Microbiology, Faculty of Natural Sciences, Chukwuemeka Odumegwu Ojukwu University, Uli Nigeria between September, 2014 and March, 2017.

Methodology: A laboratory scale study was carried on six composite samples of the sediment and water samples from the three studied areas using enrichment, screening, selection, molecular, growth effect and substrate specificity techniques.

Results: The findings revealed that screening and selection for the indigenous bacterial isolates from the three studied areas resulted in the isolation of nine out of forty eight (9/48) of the potent strains representing 18.75 % of the total isolates with significant ($P = .05$) multiple degradation and resistance potentials but with different efficiencies on xylene, anthracene and pyrene, other petroleum products and heavy metals at 50 – 300 pm and 1 %. All the nine potent strains were fully characterized molecularly and phylogenetically and belong to the genera: *Providencia*, *Alcaligenes*, *Brevundimonas*, *Myroides*, *Serratia*, and *Bacillus*.

Conclusion: Thus, these selected potent bacterial strains could significantly contribute in the development of a cost-effective bioremediation process on aromatic hydrocarbons and heavy metals contaminated environments in Nigeria.

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Keywords: Petroleum products, heavy metals, marine bacteria, multiple degradation and resistance, Niger Delta

1. INTRODUCTION

The world marine ecosystem has been studied extensively since the second half of the last century. The marine environment is subject to contamination by organic pollutants from a variety of sources. Organic contamination results from uncontrolled releases from manufacturing and refining installations, spillages during transportation, direct discharge from effluent treatment plants and run-off from terrestrial sources [1]. Environmental pollutants such as petroleum hydrocarbons, heavy metals and pesticides have been known to have direct toxic effects when released into the aquatic environment. There is a direct link between surface water and sediment contamination. Accumulated heavy metals or organic pollutants in sediment could be released back into the water with deleterious effects on human health [2].

The impact of these wastes in the Niger Delta ecosystems of Nigeria is an obvious environmental concern particularly with regards to persistence and ecotoxicity. Soil and ground water contamination by crude oil are becoming increasingly sensitive issues in Nigeria, since most of her potable water supply is derived from shallow and unconfined aquifers. It is therefore important to assess all remediation options on the basis of their ability to remove organic contaminants successfully. This is because most of these product especially the polycyclic aromatic hydrocarbons (PAHs), benzene, toluene, ethylbenzene and xylene (BTEX) and heavy metals are toxic, mutagenic and carcinogenic [3]. The physical and chemical methods like volatilization, photooxidation, chemical oxidation, and bioaccumulation are rarely successful in rapid removal and cleaning up and also these methods are not safe and cost effective when compared to microbial bioremediation [4]. A better way is to use enhanced biodegradation.

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62 Extensive studies have been done on the biodegradation of isolated bacteria from the natural environment leading to isolation of some bacteria
63 which have the ability of using PAHs compounds as the sole carbon and energy source. Isolating the bacteria with necessary performance for
64 degradation of organic pollutants such as xylene, anthracene and pyrene in soil and water ecosystems can be the perfect solution for improving
65 the microbial population in areas contaminated by hydrocarbons [5]. Several species of bacterial genera *Pseudomonas*, *Serratia*, *Marinobacter*,
66 *Providencia*, *Alcaligenes*, *Salmonella*, *Nocardia*, *Mycobacterium*, *Cunninghamella*, *Rhodococcus*, *Beijerinckia*, *Lysinibacillus*, *Corynebacterium*,
67 *Diaphorobacter*, *Pseudoxanthomonas*, *Bacillus* and *Sphingomonas* have been found highly capable of degrading petroleum hydrocarbons as well
68 as heavy metals are well documented [6, 7, 8, 9,10,11, 12, 13, 14].

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70 There are many reports by several investigators on multiple degradation and resistance potentials of bacteria from soil and fresh water
71 ecosystems on petroleum pollutants and heavy metals but few reports regarding multiple degradation and resistance capabilities of bacteria from
72 marine origin of crude oil - impacted Niger Delta ecosystem are published and hence necessitated and justifies this study. This study was
73 undertaken to determine multiple degradation and resistance potentials of marine bacteria isolated from Rivers State, Nigeria on petroleum
74 pollutants and heavy metals.

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76 **2. MATERIALS AND METHODS**

77 **2.1 Description of the Sampling Sites**

78 The studied areas were Abonema Wharf Water Front (Plate 1) in Akuku-Toru Local Government Area, Nembe Water-side (Plate 2) in Port
79 Harcourt Local Government Area and Onne Light Flow Terminal Seaport (Plate 3) located in Eleme Local Government Area of Rivers State.
80 Abonema town is 53 km and Abonema Wharf Water Front is 3 - 5 km from Port Harcourt capital city; Nembe water side is located within Port
81 Harcourt capital city of Rivers State, while Onne Light Flow Terminal is about 35 km east from Port Harcourt capital city of Rivers State and 7 km

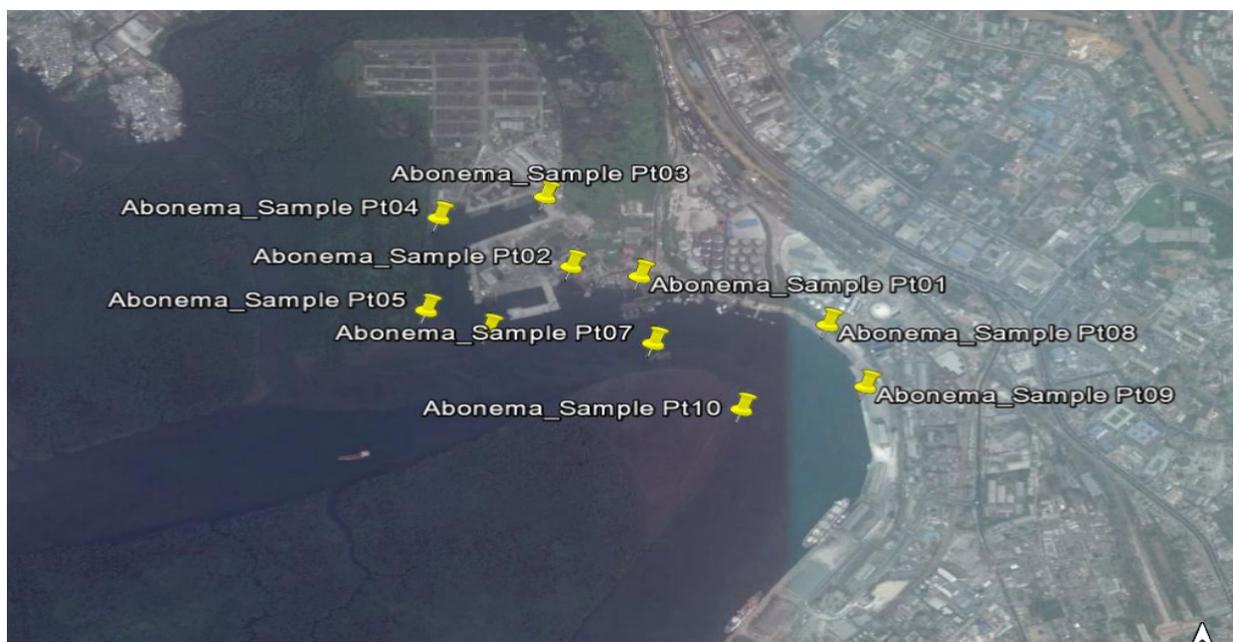
82 from Onne town. These sites were geo - referenced using Handheld Global Positioning System (GPS) GPSMAP 76 sc with the coordinates
83 obtained from the sampling points or positions Abonema Wharf Water Front, Nembe Water-side and Onne Light Flow Terminal Seaport were
84 located between latitude 4°46'15.82"N to latitude 4°46'38.01"N and longitude 7°0'0.54"E to longitude 7°0'34.82"E with average elevation of 4.1 m ,
85 latitude 4°45'8.72"N to latitude 4°45'26.42"N and longitude 7°1'11.37"E to longitude 7° 2'14.54"E with average elevation of 2.7 m and latitude
86 4°41'32.58"N and 4°41'58.18"N and longitude 7°9'26.34"E and 7°10'48.82"E with average elevation of 2.3 m, respectively. These water - ways are
87 subjected to human - induced pressures resulting from urbanization, industrialization and intensive navigation. Abonema Wharf Water Front
88 community is a popular and busy commercial but dangerous jetty area close to Port Harcourt city inhabiting tens of thousands of different families
89 living close to petroleum tank farms and tankers queue up daily to load refined petroleum products. Nembe Waterside is situated very close to
90 Creek road market, Port Harcourt, Nigeria. It shares boundary with Bayelsa and links Port Harcourt city with Bonny Island where most of the oil
91 installations in Rivers State are. It also links the Island directly with the Atlantic ocean through which crude oil is exported by massive oil tankers
92 [15]. Onne Light Flow Terminal Seaport is a port of Nigeria and the largest oil and gas free zone in the world supporting exploration and production
93 for Nigerian activities. It is situated on the Bonny River Estuary along Ogu creek and accounts for over 65 % of the export cargo through the
94 Nigerian Sea Port.

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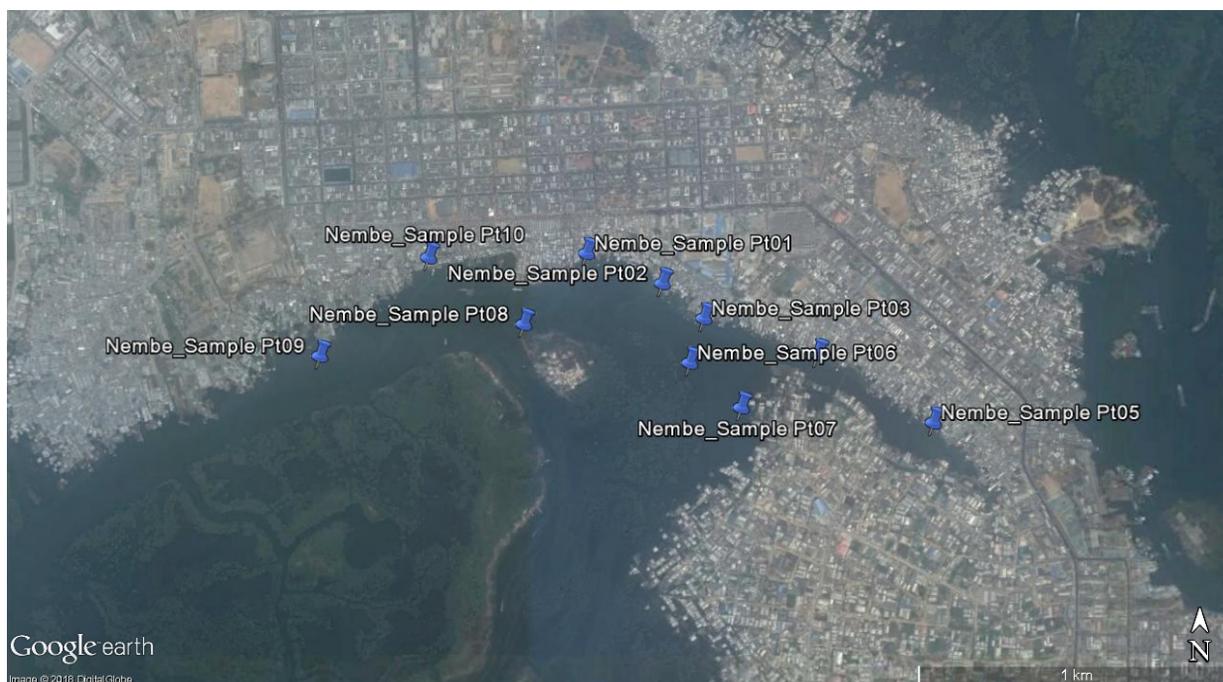
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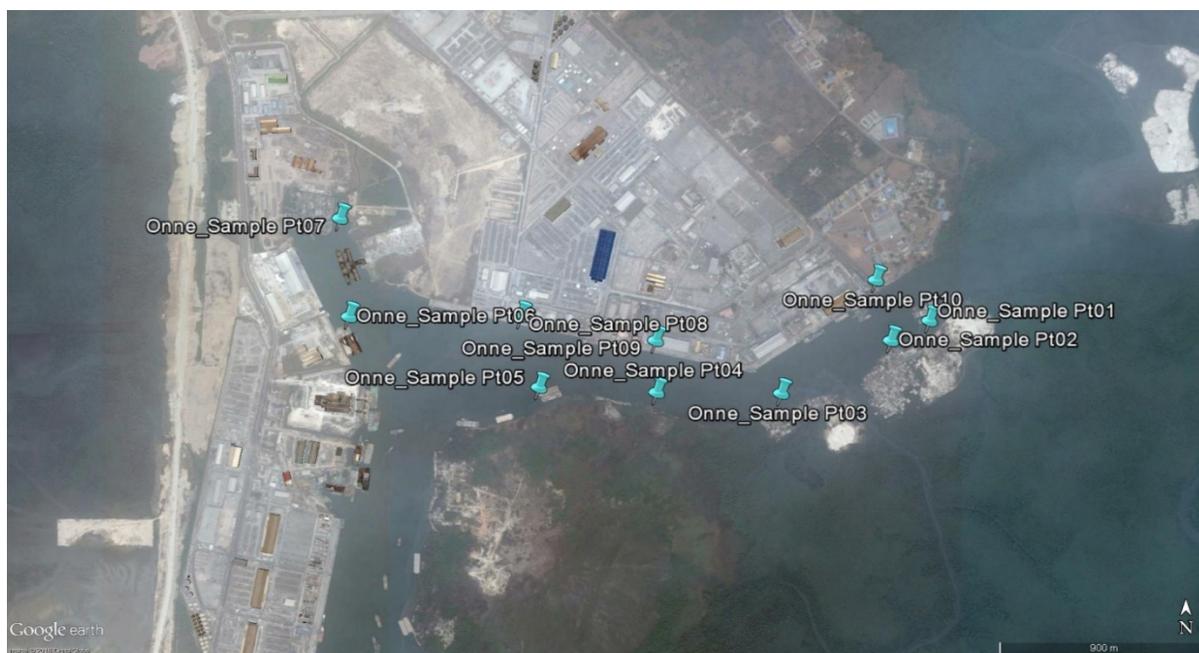
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Plate 1. Geoeye satellite image (2016) showing the Abonema sample points



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Plate 2. Geoeye satellite image (2016) showing the Nembe sample points



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Plate 3. Geoeye satellite image (2016) showing the Onne sample points

119 2.2 Sample Collection and Processing

120 Selection of sampling sites is depended on activities occurring in the region as described above (Plates 1, 2 and 3). Ten samples each of the
121 marine sediment and water were collected randomly per point of the designated ten (10) points of the three sampling sites. The samplings were
122 done once in each of the three sampling sites in September, 2014. The samples were mixed together after which a total of six composite
123 /representative sediment and water samples were taken for the analysis. The surface aerobic sediment samples were collected with a 95 %
124 ethanol - sanitized plastic spatula at 5 cm depth inside 95 % ethanol - sanitized clean, dry, leak-proof, wide mouthed plastic containers that is lined
125 with aluminum foil. The water samples were collected at the air-water interface by hand dipping the 95 % ethanol - sanitized clean, dry, leak-proof,

126 cylindrical shaped 2 L plastic containers. The containers with lids slightly opened were rinsed with the samples thrice before aseptically collecting
127 the samples. All the composite or representative sediment and water sample containers were labelled with sample type, date, time and place of
128 collection. They were placed into a sterile polythene bags in ice packed coolers to keep them under a temperature not more than 4 °C. Then
129 transported to the laboratory for microbiological analyses and stored at 4 °C in refrigerator for 12 – 24 hrs [16, 17, 18].

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131 **2.3 Enrichment, Culturing and Isolation of Aromatic Hydrocarbon Degrading Bacterial Strains**

132 The hydrocarbon degraders were isolated from sediment and water samples of the three sampling sites using modified mineral basal agar (4 g
133 K₂HPO₄, 1.0 g (NH₄)₂SO₄, 0.1 g MgSO₄, 1.8 g KH₂PO₄, 0.1 g FeSO₄, 0.1 g NaCl, 0.2 g CaCl₂, 15 g Agar agar and distilled water 1,000 mL at pH
134 7.00 ± 0.20) enriched with xylene, anthracene and pyrene as sole carbon and energy source. The medium was sterilized by autoclaving at 121 °C
135 and 15 psi for 15 minutes. Thereafter, 0.2 mL acetone solution containing 0.1 % w/v of the selected hydrocarbons (xylene, anthracene and
136 pyrene) were aseptically pipetted and uniformly spread on the agar surface of the pre - dried Petri dish plates. The acetone was allowed to
137 evaporate under sterile condition and 0.1 mL aliquots of the 10⁻³ dilutions were plated on the solidified media with a glass spreader. The spreader
138 was sterilized after each successive spreading by dipping it in 70 % ethanol and then passing it through flame of a Bunsen burner. The inoculated
139 plates were sealed using adhesive tape and foil to prevent contamination and photolysis and later placed in black polythene bags, and then
140 incubated in the dark at 28.00 ± 0.20 °C for 14 days [18, 19].

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142 **2.4 Purification and Maintenance of Cultures**

143 Colonies that developed on hydrocarbon - coated plates were replicated onto fresh hydrocarbon - coated agar plates and incubated for 14 days.
144 Isolates that grew on these plates were selected as xylene, anthracene and pyrene degraders and sub - cultured on Bijou bottles where they are
145 preserved at 4 °C in refrigerator [18].

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147 **2.5 Screening and Selection Test**

148 In order to screen and select the best and strongest degrading strains, growth of the different organisms were tested by growing 5 mL of each
149 desired isolates in large test tubes containing 25 mL of the modified mineral basal medium with 100 mg /L of xylene, anthracene and pyrene
150 hydrocarbons which were dissolved in acetone and added to each tube after autoclaving. Thereafter, the test tubes were incubated at room
151 temperature (28.00 ± 2.00 °C) for five days. Bacteria that started growing fast with high turbidity in the vicinity of the medium containing aromatic
152 compounds measured at 600 nm using a UV - VIS spectrophotometer (Astell, UV - Vis Grating, 752 W) were selected as the candidate of xylene,
153 anthracene and pyrene degrading bacteria [5,18,20].

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155 **2.6 Characterization and Identification of Selected Hydrocarbon Utilizing Bacterial Isolates**

156 **2.6.1 Molecular characteristics**

157 **2.6.1.1 Identification of bacteria**

158 Further identification was carried out using the Gram - reaction test and molecular techniques. The Gram - reaction test was first used to ascertain
159 the morphological characterisation of the colonies before proceeding to do molecular identification which includes DNA extraction, Polymerase
160 chain reaction (PCR), gel electrophoresis, sequencing and blasting [21].

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162 **2.6.1.2 Genomic DNA extraction of the bacterial isolates**

163 DNA extraction was conducted using conventional method of Cetyltrimethyl Ammonium Bromide (CTAB) protocol in sterile Eppendorf tubes.
164 About 10 mL of the pure cultures from nutrient broth was vortexed and 1.5 mL of it was transferred into 2 mL Eppendorf tubes and centrifuged with
165 a microcentrifuge (Eppendorf Minispin plus, 12 x 1.5/2.0 mL) at 14,000 rpm for 5 minutes. The supernatant was discarded to recover the pellets,

166 which was then resuspended in a solution containing 567 μL of tris ethylene diamine tetraacetic acid buffer (tris EDTA or TE buffer), 30 μL of 10 %
167 sodium dodecyl sulphate (SDS) and 3 μL of proteinase K (20 mg /m L) and was incubated in Accu block digital dry bath incubator (Labnet
168 International, USA) at 65 °C for 1 hr. Then 180 μL of 5 M NaCl and 80 μL of 10 % CTAB solutions were added to the mixture and incubated for 10
169 minutes at 65 °C. After which an equal volumes (400 mL) of phenol and chloroform was added to each tube and centrifuged at 14, 000 rpm for 15
170 minutes and then 300 μL of the supernatant was transferred into new sterile Eppendorf tubes and the DNA was precipitated by adding 0.6 mL cold
171 isopropanol to each tube. The precipitate was collected by spinning the tube in a centrifuge at 14, 000 rpm for 15 minutes and the supernatant was
172 discarded. Then 200 μL of freshly prepared 70 % ethanol was added to the tube to wash DNA pellets by spinning at 14, 000 rpm for 10 minutes.
173 The supernatant was carefully removed to air - dry the DNA pellets and 100 μL of TE buffer was added to the dried DNA pellets and incubated at
174 37 °C for 60 minutes to dissolve the DNA pellets. Then 1 μL of RNAase was added to the tube and incubated at 37 °C for 60 minutes. The DNA
175 was separated electrophoretically with 1 % agarose gel stained with 0.1 μg /mL ethidium bromide running at 80 V for 60 minutes using tris acetate
176 EDTA (TAE) electrophoresis buffer. The DNA was visualised by UV fluorescence to determine the success of the extraction process [21].

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178 **2.6.1.3 Polymerase chain reaction (PCR) and sequencing of the extracted DNA**

179 The master mix aliquot for the PCR was dispensed into individual PCR tube and the different DNA samples were added to each tubes. The
180 negative control was used to check for contamination in the master mix. The PCR reagents in each tube amounted to 50 μL containing: buffer (5
181 μL), MgCl_2 (1.5 μL), primer 1 (forward 16S - P1 PCR 5'AGAGTTTGATCCTGGCTCAG3') (2 μL), primer 2 (reverse 16S - P2 PCR
182 5'AAGGAGGTGATCCAGCCGCA3') (2 μL), dNTP mix (1 μL), Dream Taq (0.25 μL), sterile sabax water (35.25 μL) and DNA samples (3 μL). The
183 PCR reactions was performed using MJ Mini thermal cycler (Bio-Rad, Hercules, CA, USA). The cycling conditions were set at (a) initial
184 denaturation 10 minutes at 95 °C for 1 cycle. (b) Denaturation at 95 °C for 30 seconds, (c) Annealing cycling at 94 °C for 30 seconds, (d)
185 Elongation at 54 °C for 2 mins. All steps in denaturation, annealing and elongation was for 35 cycles and (e) final elongation 10 mins at 72 °C for 1

186 cycle. The reaction was held at 4 °C for 1 hr in the thermal cycler. The PCR products was separated electrophoretically with 1 % agarose gel
187 stained with 0.1 µg /mL ethidium bromide running at 80 V for 60 minutes in 100 mL of 1x TAE electrophoresis buffer. The PCR products were
188 visualised by UV fluorescence to determine the size of the amplified bands. Then the PCR products (20 µL each) were cleaned up later using 160
189 µL of 13 % polyethylene glycol (PEG) 8000, 20 µL of 5 M NaCl solution and 200 µL of 70 % ethanol ethanol. The cleaned PCR products were sent
190 for sequencing and was conducted using the automated DNA sequencer (Perkin-Elmer) which was carried out according to the manufacturers'
191 instruction. This was done at the Forestry and Agricultural Biotechnology Institute (FABI) Sequencing Facility, University of Pretoria, South Africa
192 [21].

193 **3.6.1.4 Blasting and phylogenetic correctional analyses**

194 The basic local alignment search tool (BLAST) of DNA sequences was performed by editing the sequences of the 16S rRNA region obtained
195 using BioEdit software. The edited sequences were copied in a FASTA format form. Then, blasting was done on National Centre for Biotechnology
196 Information (NCBI) website. Homologies of the 16S rRNA sequences were checked and compared with the sequences of those on the database
197 (Ubani *et al.*, 2016). From the list of many identical sequences, four of the sequences were selected for each bacterium from the GenBank for
198 phylogenetic analyses. The bacterial sequences from the treatments were aligned using online version of MAFFT software. The phylogenetic
199 correctional analyses were done using Mega 7 software and evolutionary distance of the isolates were computed using neighbour - joining (NJ)
200 methods. The bootstrap consensus tree was inferred from 100 replicates and all positions containing gaps as well as missing nucleotide data were
201 eliminated from the data - set [10, 21, 22, 23].

203 **2.6.1.5 Gene Bank Accession Number**

204 NCBI accession numbers were assigned to the nine selected aromatic hydrocarbon degrading bacterial strains and the nucleotide sequence
205 details of 16S rRNA genes of the isolates reported in this study have been deposited in the GenBank databases since 01/12/16 under the
206 accession numbers from KY171979 - KY171987 [22, 24, 25].

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208 **2.7 Determination of Concentration Effect of Aromatics on the Growth of the Isolates**

209 To determine the effects of aromatic hydrocarbon concentrations on the growth of the isolates, precisely 100 mL of modified mineral basal medium
210 was dispensed into forty-five (45) 250 mL flasks and sterilized by autoclaving. The flasks were then divided into nine sets of seven flasks.
211 Thereafter, 50, 100, 200 and 300 ppm levels of xylene, anthracene and pyrene which were separately dissolved in acetone (as before) were
212 exposed to each isolates. The fifth, sixth and seventh flasks served as the controls for each hydrocarbons and contained no xylene, anthracene
213 and pyrene. The inoculated and control flasks were then incubated as previously described at 28.00 ± 2.00 °C for 5 days. After 5 days of
214 incubation, 5 mL sample was aseptically collected from each flask and monitored for the level of microbial growth which was indicated by increase
215 in turbidity (optical density OD) of the medium measured in triplicate determinations at 600 nm using a UV - VIS spectrophotometer (Astell UV -
216 Vis Grating, 752 W) [19].

217

218 **2.8 Substrate Specificity Test**

219 In addition to growth on xylene, anthracene and pyrene, the purified strains were also tested for growth on other petroleum products and heavy
220 metals. The petroleum products include crude oil, diesel oil, kerosene, engine oil, benzene, toluene, hexane, parabenzene, ethyl benzene, and
221 phenol while the heavy metals include include copper (II) chloride (CuCl_2), arsenic trioxide (AsO_3), lead (II) sulphate (PbSO_4), mercury (II) chloride
222 HgCl_2 , potassium dichromate ($\text{K}_2\text{Cr}_2\text{O}_7$), potassium iron cyanide (KFeCN), cadmium oxide (CdO), manganese (II) chloride (MnCl_2) and zinc
223 sulphate (ZnSO_4). Modified mineral basal medium (4 g K_2HPO_4 , 1.0 g $(\text{NH}_4)_2\text{SO}_4$, 0.1 g MgSO_4 , 1.8 g KH_2PO_4 , 0.1 g FeSO_4 , 0.1 g NaCl , 0.2 g

224 CaCl₂, 15 g Agar agar and distilled water 1,000 mL at pH 7.00 ± 0.20) containing 1 % v/v of the above petroleum products were inoculated and
225 incubated for 7 days at 28.00 ± 2.00 °C while Seawater Nutrient medium containing 300 ppm of the heavy metals were inoculated and incubated
226 for 5 days at 28.00 ± 2.00 °C. After 5 days of incubation, 5 mL samples were aseptically collected from each flask and monitored for the level of
227 microbial growth which was indicated by increase in turbidity (optical density OD) of the medium measured in triplicate determinations at 600 nm
228 using a UV - VIS spectrophotometer (Astell UV - Vis Grating, 752 W). Cultures without increase in turbidity over initial optical density and non-
229 inoculated control were scored as no growth (-) while cultures with increased turbidity significantly greater than the control (i.e. growth attenuation,
230 optical density OD reading above 0.2) were scored as growth (+) [12, 19].

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232 **2.8 Data Analysis**

233 The data were analyzed using Graph-Pad Prism statistical software version 7.00 (GraphPad software Inc. San Diego, California). All values were
234 expressed as mean ± standard deviation. Ordinary one-way analysis of variance (ANOVA) followed by post Tukey's, multiple comparison test was
235 performed on the data obtained. The results were considered statistically significant at 95 % confidence intervals ($P = .05$) [5, 23].

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237 **3. RESULTS**

238 **3.1 Isolation and Selection Test**

239 The result of the growth performance (OD₆₀₀ nm) of the aromatic hydrocarbon-degraders isolated from Abonema sampled location is presented in
240 Table 1. From the result, 13 isolates were obtained with strains XYL2, ANT4 and PYR3 having the highest significant ($P = .05$) absorbance values
241 of 0.952 ± 0.004, 0.775 ± 0.007 and 1.041 ± 0.008 on xylene, anthracene and pyrene hydrocarbons. The result of the growth performance (OD₆₀₀
242 nm) of the aromatic hydrocarbon-degraders isolated from Nembe sampled location is presented in Table 2. From the result, 17 isolates were
243 obtained with strains XYL7, ANT1 and PYR5 having the highest significant ($P = .05$) absorbance values of 1.055 ± 0.002, 0.816 ± 0.007 and 0.933

244 ± 0.007 on xylene, anthracene and pyrene hydrocarbons. The result of the growth performance (OD_{600} nm) of the aromatic hydrocarbon-degraders
245 isolated from Onne sampled location is presented in Table 3. From the result, 18 isolates were obtained with strains XYL8, ANT6 and PYR9
246 having the highest significant ($P = .05$) absorbance values of 0.741 ± 0.007 , 1.433 ± 0.013 and 0.871 ± 0.001 on xylene, anthracene and pyrene
247 hydrocarbons. On the basis of these results, strains ANT1, XYL2, PYR3, ANT4, PYR5, ANT6, XYL7, XYL8 and PYR9 were selected as the best
248 and strongest degraders of xylene, anthracene and pyrene hydrocarbons.

249

250 **3.2 Molecular Characteristics of Bacterial Isolates**

251 The result of the PCR - amplification of 16S rRNA genes of the aromatic degrading bacteria genomic DNA is shown in Plate 4. From the result, it
252 was confirmed that the PCR products obtained using universal primer 16S-P1 PCR (27F 5'-3': AGA GTT TGA TCC TGG CTC AG) and 16S-P2
253 PCR (1492R 5'-3': ACG GCT ACC TTG TTA CGA CTT) have the molecular weight of 1 kbp visualized by UV fluorescence under agarose gel
254 electrophoresis. The result of the percentage similarity and Gen bank accession numbers of 16S rRNA sequences of the closest relative for the
255 aromatic degrading bacterial isolates is presented in Table 4. From the result, *Alcaligenes faecalis* was the most blasted bacteria with high
256 similarity (98 - 99 %) followed by *Providencia* spp. (95 - 97 %), *Brevundimonas diminuta* (100 %), *Myroides odoratus* (90 %), *Serratia marcescens*
257 (97 %) and *Bacillus cereus* (98 %) using NCBI BLAST software. The result of neighbor-joining phylogenetic relationship among the 16S rRNA
258 sequence of the aromatic degrading bacterial isolates constructed by MEGA 7.0 is shown in Figure 1. From the result, it revealed that the nine
259 bacterial isolates (KY171979, KY171984, KY171987, KY171980, KY171982, KY171981, KY171985, KY171986 and KY171983) show the same
260 ancestry as they arise from the same node and hence they are evolutionary related to each other.

261

262 **3.3 Effect of Aromatic Hydrocarbon Concentrations**

263 The results of the growth of the selected aromatic hydrocarbon degrading bacteria on different concentrations of xylene, anthracene and pyrene
264 are shown in Figures 2, 3 and 4. From the xylene result, the isolate *Alcaligenes faecalis* PYR5 had the least growth of 0.122 ± 0.003 (OD_{600nm})
265 observed at xylene of 300 ppm while *Providencia* sp. XYL8 had the best growth of 1.661 ± 0.297 recorded when exposed to 50 ppm of xylene for
266 5 days. From the anthracene result, the isolate *Providencia vermicola* ANT1 had the least and best growth of 0.048 ± 0.003 and 1.660 ± 0.020
267 (OD_{600nm}) when exposed to anthracene at 300 ppm and 50 ppm for 5 days respectively. From the pyrene result, the isolate *Providencia* sp. XYL8
268 had the least growth of 0.123 ± 0.001 (OD_{600nm}) observed at pyrene of 300 ppm while *Alcaligenes faecalis* XYL2 had the best growth of $1.330 \pm$
269 0.002 recorded when exposed to 50 ppm of pyrene for 5 days. There were extreme significant differences among group of cell growth
270 suspensions and the concentration of hydrocarbons ($P = .05$) with very strongly significant negative correlation ($P = .05$; $r = -0.783$ to -0.980).

271

272 3.4 Substrate Specificity Test

273 The result of the growth specific test of the marine bacterial isolates on crude oil and different hydrocarbons is presented on Table 5. From the
274 result, the abilities of the isolates to degrade different hydrocarbons were found to vary. Isolate ANT1 had moderate/heavy growth on all the
275 substrates while the other isolates had heavy, moderate, poor and no growth on the hydrocarbons substrates. The result of the growth specific test
276 of the marine bacterial isolates on different heavy metals is presented on Table 6. From the result, the abilities of the isolates to resist heavy
277 metals were also found to vary. All the isolates had growth and resisted all metals except PYR3, ANT4 and ANT6 that were sensitive (–) to AsO_3
278 and CdO compounds. Isolates ANT1, XYL2, PYR3, ANT4, PYR5, ANT6, XYL7, XYL8 and PYR 9 resisted and had heavy growths (+++) on
279 $HgCl_2$, $KFeCN$, CdO, $PbSO_4$, $CuCl_2$, $K_2Cr_2O_7$, AsO_3 , $MnCl_2$ and $ZnSO_4$ metallic compounds at 300 mg/l.

280

281 4. DISCUSSION

282 In this study, a total of nine (9) isolates ANT1, XYL2, PYR3, ANT4, PYR5, ANT6, XYL7, XYL8 and PYR9 out of the 48 isolates (9/48) representing
283 18.75 % of the isolates were screened and selected as best and strongest degraders of xylene, anthracene and pyrene hydrocarbons which they
284 significantly ($P = .05$) utilize as source of carbon and energy and is indicated by absorbance values below of each isolates (Tables 1, 2 and 3).
285 This study agrees with the explanation of Mao *et al.* [26] that the most important aspect of microbial degradation of PAH is enrichment and
286 isolation of indigenous PAH degraders because the indigenous PAH degrading bacteria are already adapted to utilizing PAH. Pathak and
287 Bhatnagar [27], argued that enrichment culturing is important for the success of hydrocarbon bioremediation because the process leads to
288 selection of microorganism accustomed to hydrocarbon degradation. Esedafe *et al.* [28] reported that an occurrence of 3/41 representing 7.32 %
289 isolates from refinery effluent were capable of utilizing phenanthrene and anthracene as sole carbon and energy sources. It also indicates that
290 only these isolates had the physiological capabilities to metabolize the aromatic hydrocarbons.

291
292 Akinbankole *et al.* [11], obtained 1,500 bp PCR product from anthracene and pyrene isolates isolated from oil contaminated water and soil in
293 Malaysia. Yuliani *et al.* [20], obtained in their research 1,489 bp PCR product from phenanthrene and pyrene isolates isolated from marine area of
294 Indonesia. Isiodu *et al.* [23], reported that all the seven (7) polycyclic aromatic hydrocarbon utilizing bacterial isolates isolated from Bodo Creek brackish
295 water in Nigeria showed amplification with an amplicon size of 500 bp. Our findings were within range of their findings. The result in Table 6
296 showed that *Alcaligenes faecalis* was the most occurring organism after blasting with high sequence similarity (98 - 99 %) followed by *Providencia*
297 spp. (95 - 97 %), *Brevundimonas diminuta* (100 %), *Myroides odoratus* (90 %), *Serratia marcescens* (97 %) and *Bacillus cereus* (98 %) using
298 NCBI BLAST software. Akinbankole *et al.* [11], found out that *B. cereus* was the most blasted organism with sequence homology (99 %). The
299 phylogenetic tree depicts the evolutionary relationship among hydrocarbon metabolizing bacteria isolated in this study. The result shown in Figure
300 1 revealed that the nine bacterial isolates (KY171979, KY171984, KY171987, KY171980, KY171982, KY171981, KY171985, KY171986 and
301 KY171983) show the same ancestry as they arise from the same node and hence they are evolutionary related to their relatives and their

302 nucleotide sequences of their 16S rRNA genes have been deposited in Genbank database since first of December two thousand and sixteen
303 (01/12/16). They belong to the genera: *Providencia*, *Alcaligenes*, *Brevundimonas*, *Myroides*, *Serratia*, and *Bacillus*; and families of:
304 Enterobacteriaceae, Alcaliginaceae, Caulobacteraceae, Flavobacteriaceae, and Bacillaceae; and phyla of: Proteobacteria, Bacteroidetes and
305 Firmicutes which members have been implicated in petroleum and aromatic hydrocarbon biodegradation by several authors [7, 8, 10, 11, 13, 15,
306 18, 20, 23, 29, 30, 31, 32].

307
308 The results in Figures 2, 3 and 4 showed that the utilization and degradation of these compounds resulted in increase in optical density (cell mass)
309 of the organisms; however, increase in concentration of these compounds led to decrease in optical density (cell mass) of the organisms with very
310 strongly significant negative correlation ($P = .05$; $r = - 0.783$ to $- 0.980$). It is apparent from the results that strains isolated on xylene (*Alcaligenes*
311 *faecalis* XYL2 and *Providencia* sp. XYL8) and anthracene (*Providencia vermicola* ANT1) hydrocarbons were able to grow better on the three
312 tested aromatic hydrocarbons than all the strains (*Brevundimonas diminuta* PYR3, *Alcaligenes faecalis* PYR5 and *Bacillus cereus* PYR9) that
313 were isolated on pyrene hydrocarbons. Also, all the nine strains degraded all the three aromatic hydrocarbons and grew well indicating multiple
314 biodegradation potentials but with different efficiencies hence termed multiple degraders. Also, these strains especially *Alcaligenes faecalis* XYL2
315 and *Providencia* sp. XYL8 degraded xylene and anthracene hydrocarbons equally but less than pyrene hydrocarbon in the order of degradation:
316 xylene = anthracene > pyrene, Comparatively, *Providencia* sp. XYL8, *Providencia vermicola* ANT1 and *Alcaligenes faecalis* XYL2 had better
317 degradation efficiencies than the rest of the other six (6) strains (*Brevundimonas diminuta* PYR3, *Alcaligenes faecalis* ANT4, *Alcaligenes faecalis*
318 PYR5, *Myroides odoratus* ANT6, *Serratia marcescens* XYL7 and *Bacillus cereus* PYR9). The result is in consistent with the research carried out by
319 Poornachander *et al.* [33], who reported that growth of *Bacillus cereus* CPOU13 decreased with increasing PAHs concentrations (phenanthrene,
320 anthracene and pyrene) from 10 ppm to 250 ppm in MSM. Also, similar was the work carried out by John *et al.* [18], in which they found out that
321 the growths of all the test isolates (*Alcaligenes faecalis* AFS-5, *P. putida* AFS-3 and *M. varians* AFS-2) were PAH - dependent and provide strong

322 evidence for selective PAH degradation by bacteria. The acclimation of microbial community to one substrate, may lead to the simultaneous
323 acclimation to some but not all structurally related molecules. Akinbankole *et al.* [11], isolated and identified *B.thuringiensis*, *B.megaterium* and
324 *B.cereus* in both pyrene and anthracene enriched medium and the three bacteria have the metabolic adaptability of utilizing low and high
325 molecular weight PAH. Bahobail *et al.* [14] reported that three isolates *Pantoea agglomerans* (BDCC-TUSA-8), *Acinetobacter lwoffii* (BDCC-
326 TUSA-12) and *Bacillus thuringiensis* (BDCC-TUSA-18) showed multiple degradation potentials with remarkably fast reaction rates on n-
327 Hexadecane, phenol and phenanthrene, representing the major types of hydrocarbon pollutants. Their abilities to utilize both low and high
328 molecular weight PAHs is an indication of the possession of ring fission enzymes [34].

329
330 The result in Table 5 revealed that the abilities of the bacterial strains to degrade different hydrocarbons were found to vary. These differences
331 might be attributed to the membrane toxicity and non - possession of the necessary enzymes [35]. The result agrees with report of Fagbemi and
332 Kehinde, [8] that the ability of the bacterial hydrocarbon degraders to degrade hydrocarbons varied and *C. koseri*, *S. ficaria* and *B. coagulans* had
333 moderate/strong growth on crude oil which is similar in this study but differ in species (*Serratia marcescens* XYL7 and *Bacillus cereus* PYR9).
334 John and Okpokwasili, [19] report that nitrifying bacteria are excellent degrader of crude oil. Also, the result in Table 6 revealed that all the isolates
335 had growth and resisted all metals except PYR3, ANT4 and ANT6 that were sensitive (-) to AsO₃ and CdO compounds. The result agrees with the
336 report of Jaysanker *et al.* [12], that single bacterial strains can be resistant to many metals and that the multi - metal resistant bacteria highly
337 resistant to mercury possess the genetic components for dealing with many toxic metal ions. These isolates are of interest for molecular
338 characterization of mechanisms for resistance to multiple metals and hold promise for bioremediation of toxic heavy metals, including in
339 environments that are contaminated by several metals. Also, Athar *et al.* [13], published that the aromatic hydrocarbon degrading bacterial isolates
340 were capable of degrading a variety of different hydrocarbons with its ability to grow in different metals stress environment. It was observed that all

341 the bacteria sensitive to the metal compounds (PbNO₃, ZnCl₂, COCl₂, CdCl₂, K₂Cr₂O₇, Hg and NiCl₂) have shown sensitivity even at lowest used
 342 concentration (50 mg /mL), whereas the resistant bacteria showed resistance to the highest concentration (150 mg /mL).

343
 344 Table 1. Growth performance of the aromatic hydrocarbon - degraders isolated from Abonema sampled location

Isolate	Optical density (OD ₆₀₀ nm)		
	Xylene	Anthracene	Pyrene
AB1	0.657 ± 0.008 ^a	0.657 ± 0.001 ^a	0.580 ± 0.003 ^a
ANT4*	0.676 ± 0.004 ^a	0.775 ± 0.007 ^a	0.822 ± 0.002 ^a
AB3	0.701 ± 0.021 ^a	0.467 ± 0.029 ^a	0.666 ± 0.003 ^a
AB4	0.715 ± 0.004 ^a	0.598 ± 0.024 ^a	0.841 ± 0.001 ^a
PYR3*	0.598 ± 0.005 ^a	0.511 ± 0.003 ^a	1.041 ± 0.008 ^a
AB6	0.641 ± 0.001 ^a	0.494 ± 0.002 ^a	0.653 ± 0.001 ^a
AB7	0.687 ± 0.004 ^a	0.690 ± 0.002 ^a	0.803 ± 0.016 ^a
AB8	0.618 ± 0.001 ^a	0.638 ± 0.007 ^a	0.782 ± 0.003 ^a
AB9	0.457 ± 0.002 ^a	0.475 ± 0.001 ^a	0.573 ± 0.004 ^a
XYL2*	0.952 ± 0.004 ^a	0.312 ± 0.002 ^a	0.838 ± 0.021 ^a
AB11	0.793 ± 0.014 ^a	0.495 ± 0.002 ^a	0.970 ± 0.003 ^a
AB12	0.647 ± 0.002 ^a	0.446 ± 0.001 ^a	0.621 ± 0.005 ^a
AB13	0.328 ± 0.001 ^a	0.415 ± 0.001 ^a	0.451 ± 0.001 ^a

* = Isolates with highest degradability; values are mean ± standard deviation of triplicate determination; superscript a = Non-significant difference;

F (12, 24) = 2.09; P = .06

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 349 Table 2. Growth performance of the aromatic hydrocarbon - degraders isolated from Nembe sampled location

Isolate	Optical density (OD ₆₀₀ nm)		
	Xylene	Anthracene	Pyrene

NW1	0.885 ± 0.003 ^a	0.236 ± 0.005 ^a	0.708 ± 0.008 ^a
PYR5*	0.710 ± 0.003 ^a	0.216 ± 0.005 ^a	0.933 ± 0.007 ^a
NW3	0.466 ± 0.007 ^a	0.201 ± 0.000 ^a	0.806 ± 0.004 ^a
NW4	0.893 ± 0.002 ^a	0.356 ± 0.008 ^a	0.827 ± 0.008 ^a
NW5	0.750 ± 0.004 ^a	0.132 ± 0.005 ^a	0.767 ± 0.008 ^a
NW6	0.644 ± 0.004 ^a	0.246 ± 0.004 ^a	0.724 ± 0.008 ^a
NW7	0.561 ± 0.003 ^a	0.193 ± 0.005 ^a	0.808 ± 0.001 ^a
NW8	0.628 ± 0.008 ^a	0.472 ± 0.001 ^a	0.826 ± 0.008 ^a
XYL7*	1.055 ± 0.002 ^a	0.588 ± 0.005 ^a	0.927 ± 0.001 ^a
NW10	0.809 ± 0.002 ^a	0.785 ± 0.002 ^a	0.881 ± 0.004 ^a
NW11	0.826 ± 0.001 ^a	0.444 ± 0.002 ^a	0.891 ± 0.001 ^a
NW12	0.625 ± 0.005 ^a	0.563 ± 0.001 ^a	0.728 ± 0.006 ^a
NW13	0.374 ± 0.008 ^a	0.775 ± 0.001 ^a	0.760 ± 0.001 ^a
NW14	0.701 ± 0.001 ^a	0.622 ± 0.003 ^a	0.788 ± 0.007 ^a
NW15	0.705 ± 0.008 ^a	0.529 ± 0.004 ^a	0.830 ± 0.002 ^a
NW16	0.769 ± 0.002 ^a	0.380 ± 0.001 ^a	0.822 ± 0.001 ^a
ANT1*	0.804 ± 0.003 ^a	0.816 ± 0.007 ^a	0.583 ± 0.001 ^a

* = Isolates with highest degradability; values are mean ± Standard deviation of triplicate determination; superscript a = Non-significant difference;
F (16, 32) = 1.08; *P* = .410.

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Table 3. Growth performance of the aromatic hydrocarbon - degraders isolated from Onne sampled location

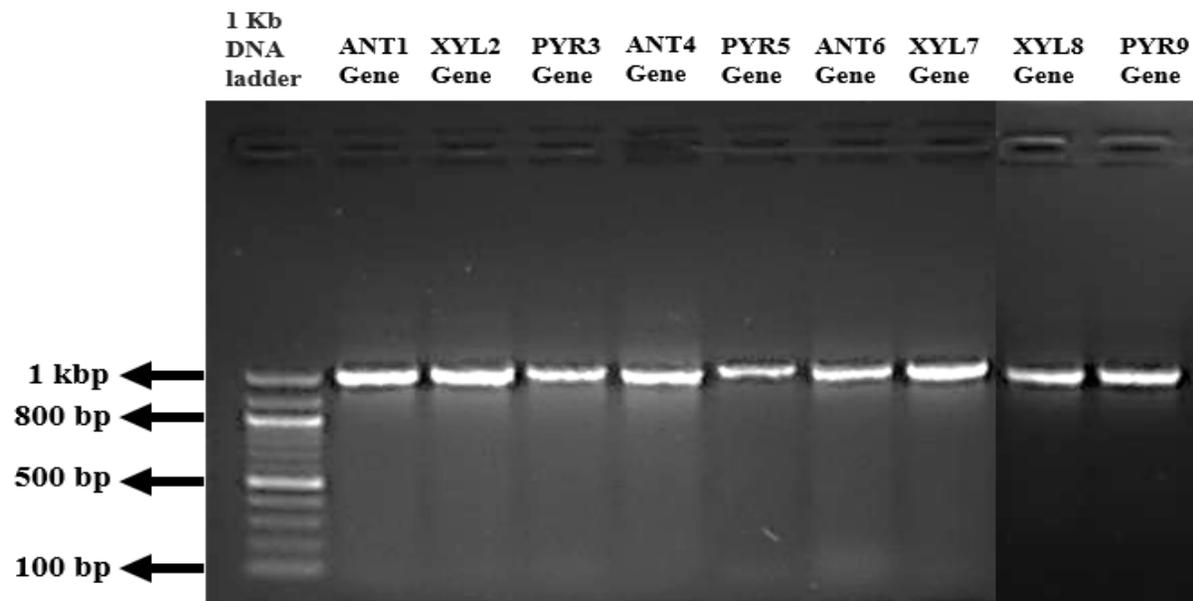
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Isolate	Optical density (OD ₆₀₀ nm)			365
	Xylene	Anthracene	Pyrene	366 367
ON1	0.721 ± 0.001 ^a	0.884 ± 0.007 ^a	0.500 ± 0.001 ^a	368
ON2	0.204 ± 0.001 ^a	0.660 ± 0.011 ^a	0.454 ± 0.001 ^a	369
ON3	0.473 ± 0.003 ^a	0.476 ± 0.036 ^a	0.561 ± 0.013 ^a	370
ON4	0.207 ± 0.001 ^a	0.766 ± 0.001 ^a	0.565 ± 0.033 ^a	371
ON5	0.477 ± 0.002 ^a	0.457 ± 0.001 ^a	0.378 ± 0.005 ^a	372
ON6	0.409 ± 0.005 ^a	0.489 ± 0.100 ^a	0.562 ± 0.021 ^a	373
ON7	0.251 ± 0.003 ^a	0.428 ± 0.014 ^a	0.728 ± 0.001 ^a	374
ON8	0.111 ± 0.005 ^a	0.429 ± 0.014 ^a	0.425 ± 0.021 ^a	375
ON9	0.463 ± 0.008 ^a	0.357 ± 0.011 ^a	0.281 ± 0.006 ^a	376
PYR9*	0.106 ± 0.001 ^a	0.335 ± 0.001 ^a	0.871 ± 0.001 ^a	377
ON11	0.700 ± 0.001 ^a	0.901 ± 0.005 ^a	0.417 ± 0.002 ^a	378
ANT6*	0.511 ± 0.006 ^a	1.433 ± 0.013 ^a	0.568 ± 0.009 ^a	379
ON13	0.273 ± 0.002 ^a	0.386 ± 0.002 ^a	0.527 ± 0.001 ^a	380
ON14	0.278 ± 0.005 ^a	0.553 ± 0.022 ^a	0.684 ± 0.003 ^a	381
ON15	0.291 ± 0.003 ^a	0.748 ± 0.009 ^a	0.522 ± 0.010 ^a	382
ON16	0.662 ± 0.001 ^a	0.919 ± 0.002 ^a	0.494 ± 0.002 ^a	383
XYL8*	0.741 ± 0.007 ^a	0.510 ± 0.013 ^a	0.602 ± 0.004 ^a	
ON18	0.354 ± 0.002 ^a	1.004 ± 0.001 ^a	0.478 ± 0.001 ^a	

* = Isolates with highest degradability; values are mean ± standard deviation of triplicate determination; superscript a = Non-significant mean;

F (17, 34) = 1.19; P = .320.

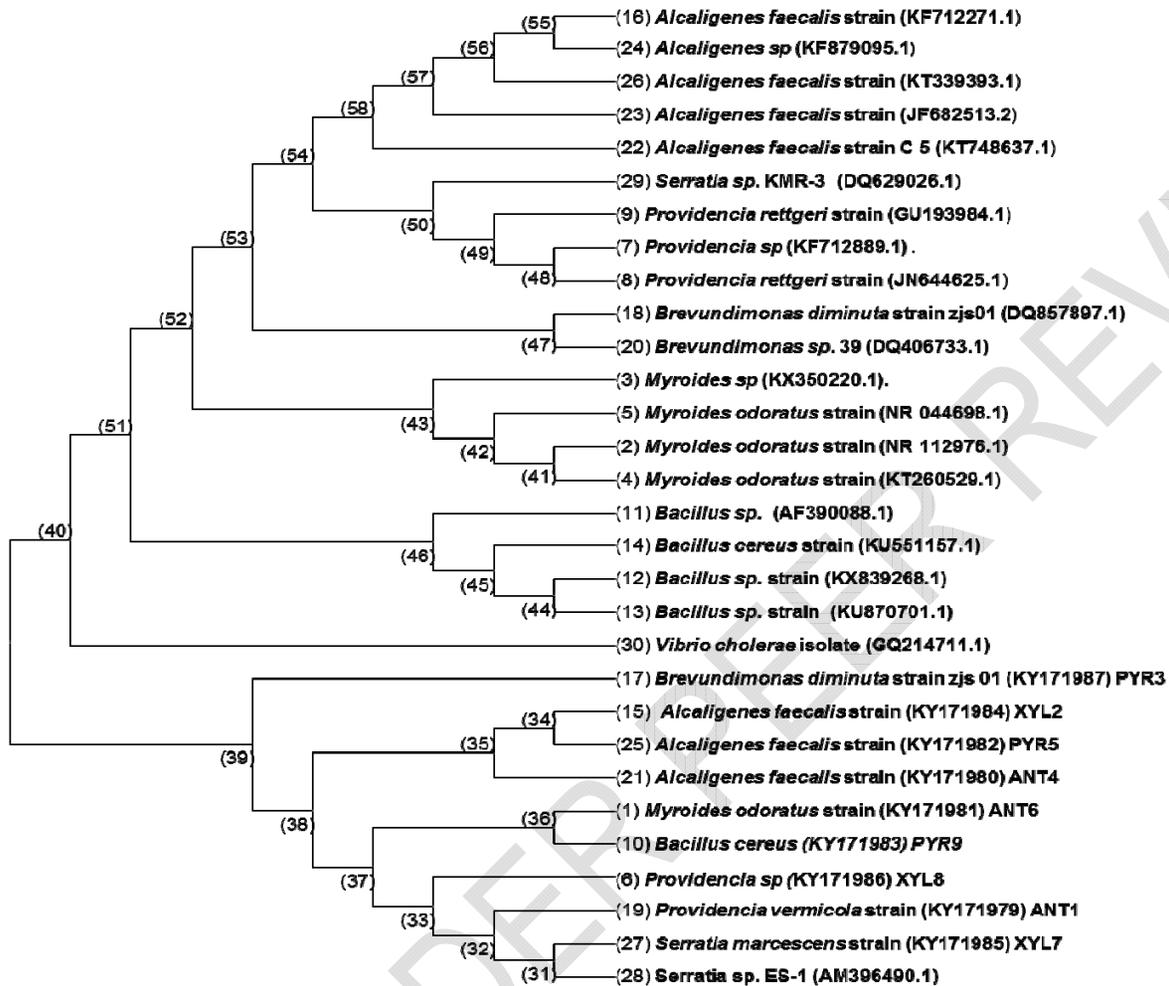
UNDER REVIEW



384
385 Plate 4. PCR - amplification of 16S rRNA genes of the aromatic degrading bacteria genomic DNA

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387 ANT1 - *Providencia vermicola*, XYL2 - *Alcaligenes faecalis*, PYR3 - *Brevundimonas diminuta*, ANT4 - *Alcaligenes faecalis*, PYR5 - *Alcaligenes faecalis*, ANT6 - *Myroides odoratus*, XYL7 - *Serratia marcescens*, XYL8 -
388 *Providencia* sp. and PYR9 - *Bacillus cereus*.

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Fig. 1. Neighbor-joining phylogenetic relationship among the 16S rRNA sequence of the aromatic degrading bacterial isolates constructed by MEGA 7.0.

Numbers at the nodes indicate bootstrap support (%) based on 1000 replicates. The sum of branch length = 0.93646865 using *p* - distance method involving 9 nucleotide sequences with total of 303 positions. GenBank accession numbers are given in parentheses.

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Table 4. Percentage similarity and Gen Bank accession numbers of 16S rRNA sequences of the closest relative for the aromatic degrading bacterial isolates

Isolate code	Closest relative in Gen bank	Max score	Total score	Query coverage	E. value	Max identity	Accession Number
ANT1 Nembe	<i>Providencia vermicola</i> strain MTCC 5578	544	544	93%	6e-151	95%	KY171979 ⁴⁰⁹
XYL2 Abonema	<i>Alcaligenes faecalis</i> strain MOR02	1559	4679	99%	0.0	99%	KY171984 ⁴¹⁰
PYR3 Abonema	<i>Brevundimonas diminuta</i> strain zjs 01	1489	1489	97%	0.0	100%	KY171987 ⁴¹¹
ANT4 Abonema	<i>Alcaligenes faecalis</i> strain MOR02	1594	1594	99%	0.0	99%	KY171980 ⁴¹²
PYR5 Nembe	<i>Alcaligenes faecalis</i> strain MOR02	1537	4612	98%	0.0	98%	KY171982 ⁴¹³
ANT6 Onne	<i>Myroides odoratus</i> strain D25T	1194	1194	95%	0.0	90%	KY171981 ⁴¹⁴
XYL7 Nembe	<i>Serratia marcescens</i> strain SM6	1476	1476	98%	0.0	97%	KY171985 ⁴¹⁵
XYL8 Onne	<i>Providencia</i> sp. strain X1	1491	1491	98%	0.0	97%	KY171986 ⁴¹⁶
PYR9 Onne	<i>Bacillus cereus</i> strain B4	1543	16940	98%	0.0	98%	KY171983 ⁴¹⁷

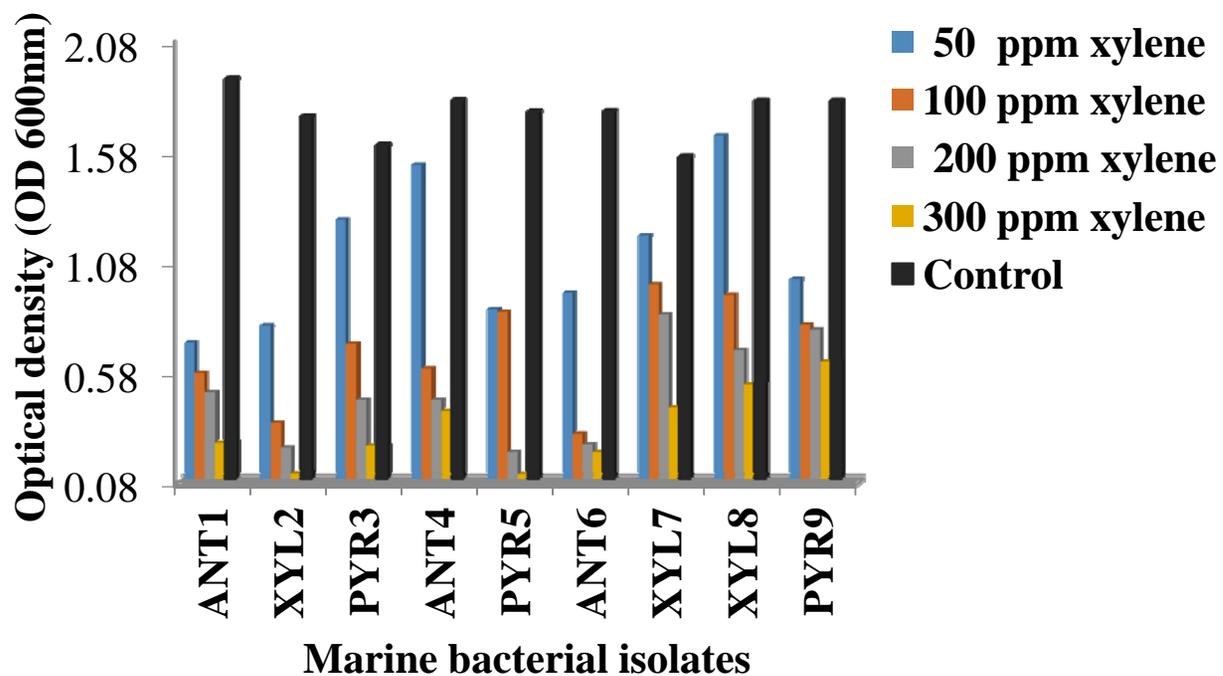


Fig. 2. Growth of the selected aromatic hydrocarbon degrading bacteria on different concentrations of xylene

PPM = Part Per Million; XYL = Xylene; ANT = Anthracene; PYR = Pyrene; ANT1 - *Providencia vermicola*, XYL2 - *Alcaligenes faecalis*, PYR3 - *Brevundimonas diminuta*, ANT4 - *Alcaligenes faecalis*, PYR5 - *Alcaligenes faecalis*, ANT6 - *Myroides odoratus*, XYL7 - *Serratia marcescens*, XYL8 - *Providencia sp.* and PYR9 - *Bacillus cereus*; $F(9, 40) = 5.75$; $P < 0.0001$; $R^2 = 0.614 - 0.944$.

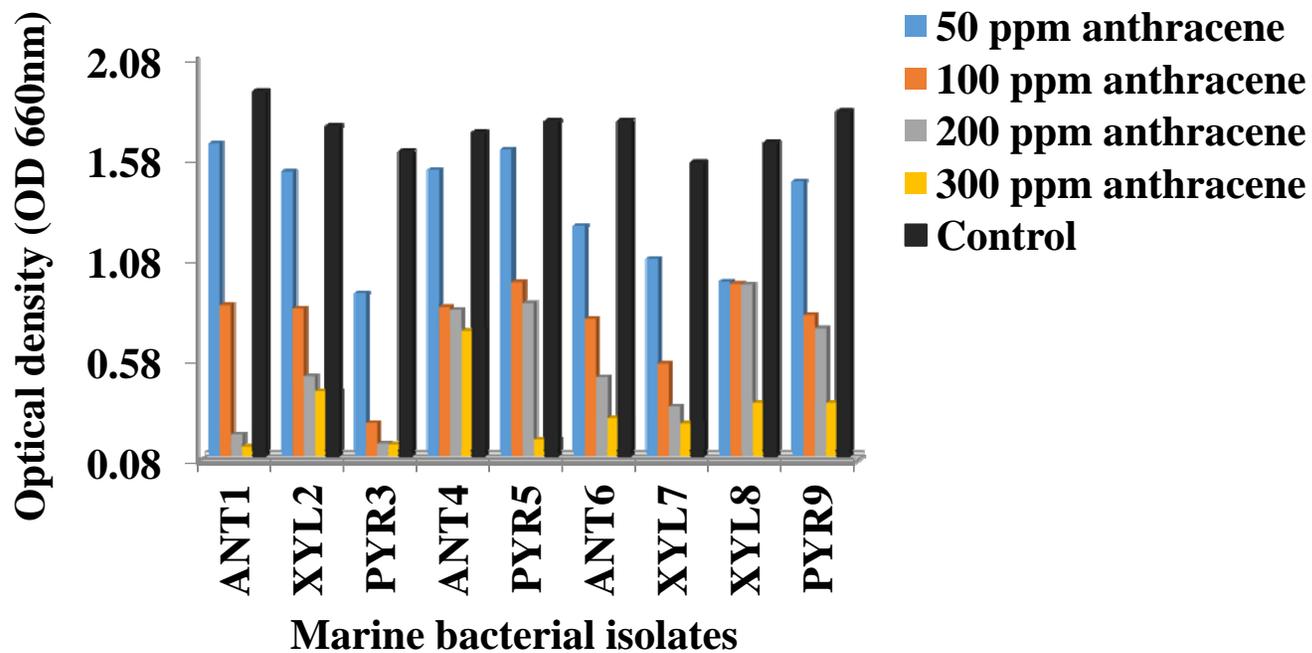
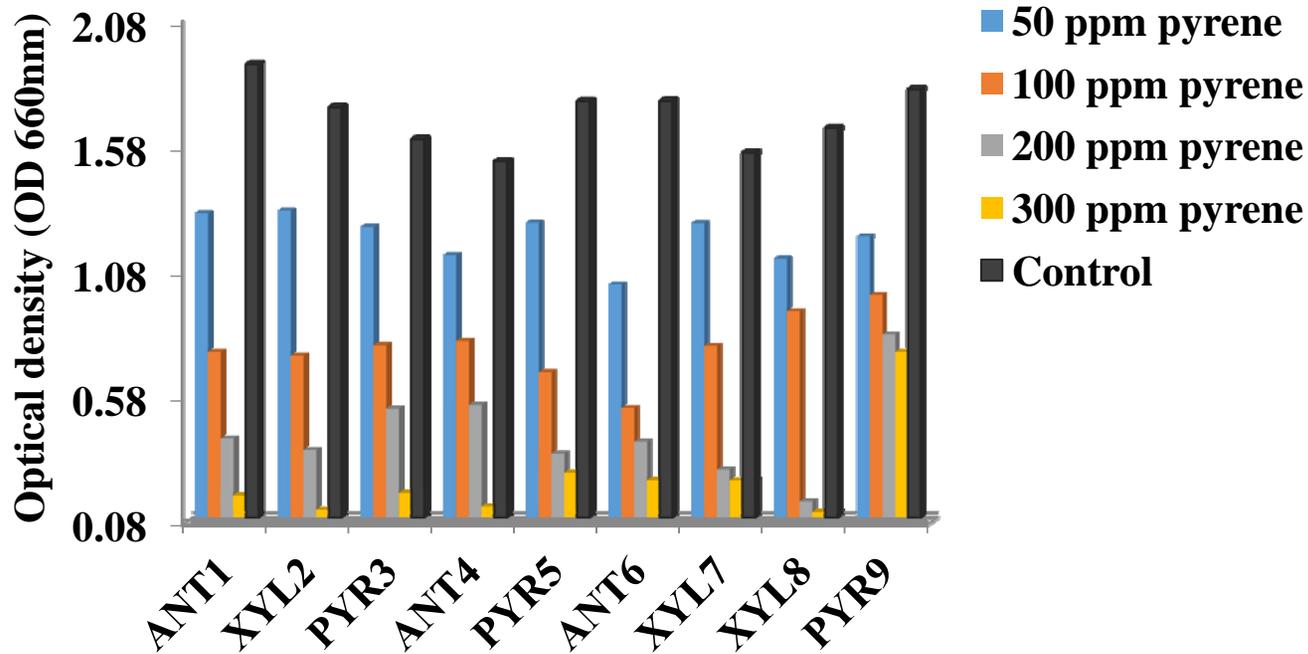


Fig.3. Growth of the selected aromatic hydrocarbon degrading bacteria on different concentrations of anthracene

PPM = Part Per Million; XYL = Xylene; ANT = Anthracene; PYR = Pyrene; ANT1 - *Providencia vermicola*, XYL2 - *Alcaligenes faecalis*, PYR3 - *Brevundimonas diminuta*, ANT4 - *Alcaligenes faecalis*, PYR5 - *Alcaligenes faecalis*, ANT6 - *Myroides odoratus*, XYL7 - *Serratia marcescens*, XYL8 - *Providencia sp.* and PYR9 - *Bacillus cereus*; $F(9, 40) = 5.74$; $P < 0.0001$; $R^2 = 0.676 - 0.932$.



Marine bacterial isolates

Fig. 4. Growth of the selected aromatic hydrocarbon degrading bacteria on different concentrations of pyrene
 PPM = Part Per Million; XYL = Xylene; ANT = Anthracene; PYR = Pyrene; ANT1 - *Providencia vermicola*, XYL2 - *Alcaligenes faecalis*, PYR3 - *Brevundimonas diminuta*, ANT4 - *Alcaligenes faecalis*, PYR5 - *Alcaligenes faecalis*, ANT6 - *Myroides odoratus*, XYL7 - *Serratia marcescens*, XYL8 - *Providencia sp.* and PYR9 - *Bacillus cereus*; $F(9, 40) = 5.75$; $P < 0.0001$; $R^2 = 0.739 - 0.961$.

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424 Table 5. Growth specific test of the marine bacterial isolates on crude oil and different hydrocarbons

Substrate	Isolate
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	<i>Providencia vermicola</i> strain ANT1	<i>Alcaligenes faecalis</i> strain XYL2	<i>Brevundimonas diminuta</i> strain PYR3	<i>Alcaligenes faecalis</i> strain ANT4	<i>Alcaligenes faecalis</i> strain PYR5	<i>Myroides odoratus</i> strain ANT6	<i>Serratia marcescens</i> strain XYL7	<i>Providencia</i> sp. strain XYL8	<i>Bacillus cereus</i> strain PYR9
Crude oil	++	+++	+++	+++	++	+++	++	+	+++
Toluene	+++	+++	++	+++	-	-	-	-	-
Petrol	+++	-	++	++	+	-	-	+++	++
Petroleum ether	++	++	+	+++	-	++	+++	-	+
Engine oil	+++	++	+++	+++	++	+++	+	++	+++
Diesel	+	+	+++	-	+	+++	++	+++	+++
Hexane	+++	-	-	+++	+	+++	+++	+	+++
Kerosene	++	+	+++	++	+	+	+++	++	++
Catechol	+++	+	+	+	+	+++	+	++	+
Parabenzene	++	+	++	+++	++	++	+++	+++	+++

Growth was followed by measuring the increase of OD at 600 nm of the culture for 7 days; +++ Heavy growth: OD 600 nm > 0.2; ++ Moderate growth: OD 600 nm 0.1 ≤ 0.2

+ Poor growth: OD 600 nm 0.02 ≤ 0.; - No growth: OD 600 nm < 0.02

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428 Table 6. Growth specific test of the marine bacterial isolates on different heavy metals

Isolate	Substrate (300 mg/l)									
	CuCl ₂	AsO ₃	PbSO ₄	HgCl ₂	K ₂ Cr ₂ O ₇	KFeCN	CdO	MnCl ₂	ZnSO ₄	
<i>Providencia vermicola</i> strain ANT1	++	+	+	+++	++	++	+	+	+	
<i>Alcaligenes faecalis</i> strain XYL2	++	+	+	++	++	+++	+	+	+	
<i>Brevundimonas diminuta</i> strain PYR3	+	-	+	+	+	+	+++	+	+	
<i>Alcaligenes faecalis</i> strain ANT4	+	-	+++	+	+	+	+	+	++	
<i>Alcaligenes faecalis</i> strain PYR5	+++	+	+	++	++	++	+	+	+	

<i>Myroides odoratus</i> strain ANT6	++	-	+	++	+++	++	-	+	+
<i>Serratia marcescens</i> strain XYL7	+	+++	+	+	+	+	-	+	+
<i>Providencia</i> sp. strain XYL8	+	+	+	+	+	+	+	+++	+
<i>Bacillus cereus</i> strain PYR9	+	+	++	+	+	+	+	+	+++

Growth was followed by measuring the increase of OD at 600 nm of the culture for 5 days; +++ Heavy growth: OD 600 nm > 0.2

++ Moderate growth: OD 600 nm 0.1 ≤ 0.2; + Poor growth: OD 600 nm 0.02 ≤ 0.1; - No growth: OD 600 nm < 0.02

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438 5. CONCLUSION

439 It can be concluded that screening for indigenous bacterial isolates from the three studied areas resulted in the isolation of nine out of forty eight
440 (9/48) potent isolates with multiple degradation and resistance potentials on xylene, anthracene, pyrene, other petroleum products and heavy
441 metals. All the nine potent strains were fully characterized, identified and belong to the genera: *Providencia*, *Alcaligenes*, *Brevundimonas*,
442 *Myroides*, *Serratia*, and *Bacillus*. Together with their demonstrated multiple degradation and resistance potentials, the nine selected potent

443 bacterial strains could significantly contribute in the development of a cost-effective bioremediation process on aromatic hydrocarbons
444 contaminated environments in Nigeria.

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447 **COMPETING INTERESTS**

448 Author has declared that no competing interests exist.

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452 **REFERENCES**

- 453 1. Hassanshahian M. The effects of crude oil on marine microbial communities in sediments from the Persian Gulf and the Caspian Sea: A
454 microcosm experiment. *Internat J Adv Biol Biomed Res.* 2014;2(1):1-17.
- 455 2. Taiwo AM, Olujimi OO, Bamgbose O, Arowolo TA. Surface water quality monitoring in Nigeria: Situational analysis and future management
456 strategy, water quality monitoring and assessment. Dr. Voudouris (Ed.). *InTech.* 2012; 978-953-51-0486-5. Accessed 05 April 2017. Available
457 from: [http://www.intechopen.com/books/waterquality-monitoring-and-assessment/surface-water-quality-monitoring-in-nigeria-situational-analysis-](http://www.intechopen.com/books/waterquality-monitoring-and-assessment/surface-water-quality-monitoring-in-nigeria-situational-analysis-and-futuremanagement-strategy)
458 [and-futuremanagement- strategy.](http://www.intechopen.com/books/waterquality-monitoring-and-assessment/surface-water-quality-monitoring-in-nigeria-situational-analysis-and-futuremanagement-strategy)
- 459 3. Akpe AR, Ekundayo AO and Esumeh FI. Degradation of crude oil by bacteria: A role for plasmid-borne genes. *Glob J Sci Front Res*
460 2013;13(6):20-6.
- 461 4. Dasgupta D, Ghosh R, and Sengupta TK. Biofilm-mediated enhanced crude oil degradation by newly isolated *Pseudomonas* species. *ISRN*
462 *Biotech.* 2013;2013(250749):1-13.
- 463 5. Kafilzadeh F, Pour FH. Degradation of naphthalene, phenanthrene and pyrene by *Pseudomonas* sp. and *Corynebacterium* sp. in the landfills.
464 *Internat J Biosci.* 2012;2(9):77-84.
- 465 6. Babita K, Singh SN, Deeba F, Sharma M, Pandey V, Singh DP. Elucidation of pyrene degradation pathway in bacteria. *Adv Biores.*
466 2013;4(2):151-60.
- 467 7. Amer RA, Mapelli F, El-Gendi HM, Barbato M, Goda DA, Corsini A et al. Bacterial diversity and bioremediation potential of the highly
468 contaminated marine sediments at El-Max District (Egypt, Mediterranean Sea). *Biomed Res Internat* 2015;2015:981-9.
- 469 8. Fagbemi OK, Sanusi A. I. Chromosomal and plasmid mediated degradation of crude oil by *Bacillus coagulans*, *Citrobacter koseri* and *Serratia*
470 *ficaria* isolated from the soil. *Afr J Biotech.* 2017;16(21):1242-53.
- 471 9. Lily MK, Bahuguna A, Bhatt KK, Dangwal K. Degradation of anthracene by a novel strain *Brachybacterium paraconglomeratum* BMIT637C
472 (MTCC 9445). *Internat J Environ Sci.* 2013;3(4):1242-52.
- 473 10. Swaathy S, Kavitha V, Pravin AS, Mandal, AB, Gnanamani A. Microbial surfactant mediated degradation of anthracene in aqueous phase by
474 marine *Bacillus licheniformis* MTCC 5514. *Biotech Rep* 2014a;4:161-70.

- 475 11. Akinbankole AS, Tunung R, Tennant AM. Biochemical and molecular characterization of pyrene and anthracene metabolizing bacteria isolated
476 from oil contaminated water and soil in Malaysia. *J Appl Environ Microbiol.* 2015;3(1):25-30.
- 477 12. Jaysanker D, Ramaiah N, Bhosle NB, Garg A, Vardanyan L, Nagle VL, Fukami K. Potential of mercury resistant marine bacteria for
478 detoxification of chemicals of environmental concern. *Microb Environ.* 2007;22 (4):336-45.
- 479 13. Athar MA, Akbar A, Khan YH, Ali I, Mehmood U, Sabri AN, Hasnain S. Characterization of hydrocarbon degrading bacteria isolated from
480 hydrocarbon contaminated soil. *J Pure Appl Microbiol.* 2014;8(6):1-9.
- 481 14. Bahobail A., Gad El-Rab SMF, Amin GA. Locally isolated bacterial strains with multiple degradation potential capabilities on petroleum
482 hydrocarbon pollutants. *Adv Microbiol.* 2016;6:852-66.
- 483 15. Chikere CB, Okpokwasili GC, Ichiakor O. Characterization of hydrocarbon utilizing bacteria in tropical marine sediments. *Afr J Biotech.*
484 2009;8(11):2541-4.
- 485 16. Abu GO, Chikere BO. Cell surface properties of hydrocarbon utilizing bacterial isolates from the Port Harcourt marine environment. *Nig J*
486 *Microbiol.* 2006;20(1):809-16.
- 487 17. Gorleku MA, Carboo D, Palm LMN, Quasie WJ, Armah AK. Polycyclic aromatic hydrocarbons (PAHs) pollution in marine waters and
488 sediments at the Tema Harbour, Ghana. *Acad J Environ Sci.* 2014;2 (7):108-15.
- 489 18. John RC, Essien JP, Akpan SB, Okpokwasili GC. Polycyclic aromatic hydrocarbon degrading bacteria from aviation fuel spill site at Ibeno,
490 Nigeria. *Bull Environ Contam Toxicol.* 2012;88:1014-9.
- 491 19. John RC, Okpokwasili GC. Crude oil-degradation and plasmid profile of nitrifying bacteria isolated from oil impacted mangrove sediment in the
492 Niger Delta of Nigeria. *Bull Environ Contam Toxicol.* 2012;88:1020-26.
- 493 20. Yuliani H, Sahlan M, Hermansyah H, Wijanarko A. Selection and identification of polyaromatic hydrocarbon degrading bacteria. *World Appl Sci*
494 *J.* 2012;20(8):1133-8.
- 495 21. Ubani O, Atagana HI, Thantsha MS, Rasheed A. Identification and characterisation of oil sludge degrading bacteria isolated from compost.
496 *Arch Environ Prot.* 2016;42(2):67-77.
- 497 22. Hesham AE, Mawad AMM, Mostafa YM, Shoreit A. Biodegradation ability and catabolic genes of petroleum-degrading *Sphingomonas*
498 *koreensis* strain ASU-06 isolated from Egyptian oily soil. *BioMed Res Internat.* 2014;2014:127674.
- 499 23. Isiodu GG, Stanley HO, Ezebuoro V, Okerentugba PO. Role of plasmid-borne genes in the biodegradation of polycyclic aromatic hydrocarbons
500 (PAHs) by consortium of aerobic heterotrophic bacteria. *J Petr Environ Biotech.* 2016;7(1):264.
- 501 24. Swaathy S, Kavitha V, Pravin AS, Sekaran G, Mandal AB, Gnanamani A. (2014b). Phylogenetic framework and biosurfactant gene expression
502 analysis of marine *Bacillus* spp. of Eastern Coastal Plain of Tamil Nadu. *Internat J Bact.* 2014b;2014:860491.
- 503 25. Guo C, Dang Z, Wong Y, Tam NF. Biodegradation ability and dioxygenase genes of PAH-degrading *Sphingomonas* and *Mycobacterium* strains
504 isolated from mangrove sediments. *Internat Biodet Biodeg.* 2010;64:419-26.
- 505 26. Mao J, Luo Y, Teng Y, Li Z. Bioremediation of polycyclic aromatic hydrocarbon-contaminated soil by a bacterial consortium and associated
506 microbial community changes. *Internat Biodet Biodeg.* 2012;70:141-7.
- 507 27. Pathak H, Bhatnagar K. *Alcaligenes*-the 4T engine oil degrader. *J Biorem and Biodeg.* 2011;2(4):2-5.
- 508 28. Esedafe WK, Fagade OE, Umaru FF, Akinwotu O. Bacterial degradation of the polycyclic aromatic hydrocarbon (PAH) -fraction of refinery
509 effluent. *Internat J Environ Biorem Biodeg.* 2015;3(1):23-7.
- 510 29. Ichor T, Okerentugba PO, Okpokwasili GC. Biodegradation of total petroleum hydrocarbon by aerobic heterotrophic bacteria isolated from
511 crude oil contaminated brackish waters of Bodo Creek. *J Biorem Biodeg* 2014;5(5):236-41.
- 512 30. Irshaid FI, Jacob JH. Screening and characterization of aerobic xylene-degrading bacteria from gasoline contaminated soil sites around gas
513 stations in Northern Jordan. *J Biol Sci.* 2015;15(4):167-76.

- 514 31. Louvado A, Gomes NCM, Simões MMQ, Almeida A, Cleary DFR, Cunha A. Polycyclic aromatic hydrocarbons in deep sea sediments:
515 microbe–pollutant interactions in a remote environment. *Sci Tot Environ.* 2015;526:312-28.
- 516 32. Patowary K, Saikia RR, Kalita MC, Deka S. Degradation of polyaromatic hydrocarbons employing biosurfactant-producing *Bacillus pumilus*
517 KSS2. *Ann Microbiol.* 2015;65:225-34.
- 518 33. Poornachander RM, Anitha Y, Satyaprasad K. Abilities of *Bacillus Cereus* CPOU13 in biodegradation of polycyclic aromatic hydrocarbons
519 (PAHs). *J Pharm Chem Biol Sci.* 2016;4(1):54-64.
- 520 34. Amund OO, Ilori MO, Adebuseye SA, Musa KJ. Utilization of alicyclic compounds by soil bacteria. *Nat Sci.* 2006;4(3):65-8.
- 521 35. Igwo-Ezikpo MN, Gbenle OG, Ilori MO. Growth study on chrysene degraders isolated from polycyclic aromatic hydrocarbon polluted soils in
522 Nigeria. *Afr J Biotech.* 2006;5:823-8.
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